Pregnancy exposure to phthalates and DNA methylation in male placenta - An epigenome-wide association study

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ABSTRACT

Background: Exposure to phthalates during pregnancy may alter DNA methylation in the placenta, a crucial organ for the growth and development of the fetus.

Objectives: We studied associations between urinary concentrations of phthalate biomarkers during pregnancy and placental DNA methylation.

Methods: We measured concentrations of 11 phthalate metabolites in maternal spot urine samples collected between 22 and 29 gestational weeks in 202 pregnant women. We analyzed DNA methylation levels in placental tissue (fetal side) collected at delivery. We first investigated changes in global DNA methylation of repetitive elements *Alu* and LINE-1. We then performed an adjusted epigenome-wide association study using IlluminaHM450 BeadChips and identified differentially methylated regions (DMRs) associated with phthalate exposure.

Results: Monobenzyl phthalate concentration was inversely associated with placental methylation of *Alu* repeats. Moreover, all phthalate biomarkers except for monocarboxy-*iso*-octyl phthalate and mono(2-ethyl-5-hydroxyhexyl) phthalate were associated with at least one DMR. All but three DMRs showed increased DNA methylation with increased phthalate exposure. The largest identified DMR (22 CpGs) was positively associated with monocarboxy-*iso*-nonyl phthalate and encompassed heat shock proteins (*HSPA1A*, *HSPA1L*). The remaining DMRs encompassed transcription factors and nucleotide exchange factors, among other genes.

Conclusions: This is the first description of genome-wide modifications of placental DNA methylation in association with pregnancy exposure to phthalates. Our results suggest epigenetic mechanisms by which exposure to these compounds could affect fetal development. Of interest, four identified DMRs had been previously associated with maternal smoking, which may suggest particular sensitivity of these genomic regions to the effect of environmental contaminants.

1. Introduction

The placenta plays an important role in the paradigm of the developmental origins of health and disease (DOHaD), a concept focusing on the role of the prenatal environment in determining the development of disease later in life (Gillman 2005). Apart from transporting nutrients and waste products between mother and fetus, the placenta affects the programming of the fetal phenotype. There is growing evidence for epigenetics playing an important role in this process (Maccani and Marsit 2009; Robinson et al. 2019). Placental epigenetic mechanisms can be sensitive to environmental factors, such as chemicals, several of which have been associated with different developmental disorders, birth defects, and child health problems (Kishi and Grandjean 2020). The placental unique epigenetic landscape could also serve as a "molecular archive" of the fetal developmental environment (Heijmans et al. 2009).

Herein we focused on phthalates, a family of non-persistent chemicals abundant in the environment due to their broad spectrum of applications including in solvents, as plasticizers and additives in polyvinyl chloride plastics or personal care products (Latini 2005). Some phthalates can cross the placental barrier and their metabolites have been detected in the placental tissue (Mose et al. 2007b, 2007a). There is also growing epidemiological evidence that pregnancy exposure to several phthalates may be associated with different placental epigenetic endpoints (reviewed by Vlahos et al. 2019), DNA methylation being one of them (reviewed by Dutta et al., 2020; Strakovsky and Schantz 2018). Nevertheless, epidemiological studies on associations between phthalate exposure during pregnancy and DNA methylation marks in placenta are scarce (LaRocca et al. 2014; Zhao et al. 2015b, 2016). These studies relied on a candidate gene approach in normal or complicated pregnancies (i.e., fetal growth restriction newborns) and included two genes at most. LaRocca et al. focused on the imprinted *H19* and insulin like growth factor 2 (*IGF2*) genes (n = 179, LaRocca et al. 2014) while Zhao et al. on *IGF2* and growth-related aryl-hydrocarbon receptor repressor (*AHRR*) genes (n = 181, Zhao et al. 2016) or LINE-1 repetitive elements (n = 119, Zhao et al. 2015b). The only existing epigenome-wide study focused on 16 early terminated pregnancies, which limits the generalizability of the findings to fully developed pregnancies (Grindler et al. 2018). In the present study we hypothesized that pregnancy exposure to phthalates impacts DNA methylation profiles in placenta. Therefore, we investigated the associations between maternal concentrations of 11 phthalate metabolites and genome-wide DNA methylation in placentas collected at birth. We assessed global DNA methylation relying on repetitive elements *Alu* and LINE-1 and performed an adjusted epigenome-wide association study (EWAS) using IlluminaHM450 BeadChips. We also identified differentially methylated regions (DMRs) associated with phthalate exposure.

2. Methods

2.1. Study design and population

We relied on a subsample of 202 mother-son pairs from the French mother–child cohort EDEN (Etude des Déterminants pré et postnatals du développement et de la santé de l'Enfant) recruited between 2003 and 2006 (Heude et al. 2016). Recruitment of pregnant women at the Nancy and Poitiers University hospitals took place before their 24th week of gestation. Exclusion criteria were: maternal diabetes before pregnancy, multiple fetuses, intention to deliver outside the university hospital or to move out of the study region within the next three years, and inability to speak French. Out of the 2,002 enrolled participants, 1,301 had placental samples collected and, for 668 individuals, placental DNA methylation was assessed. Out of those, 202 women delivering a boy had phthalate metabolite concentrations assessed in urine (Botton et al. 2016; Chevrier et al. 2012) and available information on covariates (Supplementary Fig. 1). Phthalate metabolites assessment in urine was restricted to pregnancies with male fetuses in the context of a previous project focusing on their associations with male congenital malformations (Chevrier et al. 2012). The characteristics [maternal smoking status, maternal pre-pregnancy body mass index (BMI), maternal and gestational age] of the 798 mother-son pairs excluded from this study were similar to those of the included participants (results not shown).

The EDEN cohort received approval from the ethics committee (CCPPRB) of Kremlin Bicêtre and from the French data privacy institution "Commission Nationale de l'Informatique et des Libertés". Written consent was obtained from the mother for herself and for the offspring. The involvement of the Centers for Disease Control and Prevention (CDC) laboratory did not constitute engagement in human subjects' research.

2.2. Assessment of phthalate metabolites in maternal urine

Between 22 and 29 gestational weeks, women were asked to collect a spot sample of a first morning urine void at home, before a study visit. Women who did not collect their urine at home, collected a spot sample at the hospital during the visit. Urine samples were aliquoted and stored on dry ice at −80 °C before shipment to the National Center for Environmental Health laboratory at the CDC in Atlanta, Georgia, for the assessment of phthalate metabolite concentrations. Eleven phthalate metabolites were measured: monoethyl phthalate (MEP), mono-*iso*-butyl phthalate (MiBP), mono-n-butyl phthalate (MnBP), monobenzyl phthalate (MBzP), mono(3-carboxypropyl) phthalate (MCPP), mono-carboxy-*iso*-octyl phthalate (MCOP), monocarboxy-*iso*-nonyl phthalate (MCNP), and four metabolites of di-2-ethylhexyl phthalate (DEHP): mono(2-ethylhexyl) phthalate (MEHP), mono(2-ethyl-5-hydroxyhexyl) phthalate (MEHHP), mono(2-ethyl-5-oxohexyl) phthalate (MEOHP), mono(2-ethyl-5-carboxypentyl) phthalate (MECPP). Metabolite concentrations were quantified using online solid phase extraction-high performance liquid chromatography-isotope dilution-tandem mass spectrometry (Ye et al. 2005). Creatinine, a marker of urine dilution, was also measured.

2.3. Phthalate metabolite concentrations, imputation and standardization

Urinary concentrations (μg/L) below the limit of detection (LOD) were substituted by instrumental reading values. If the instrumental reading values equaled zero (i.e., indicating no signal) they were replaced by the lowest instrumental reading value specific for each phthalate and divided by $\sqrt{2}$. Concentrations were then standardized on sampling conditions (gestational age at collection, day of sampling, hour of sampling, year of sample analysis at the CDC, and duration of storage at room temperature before freezing), and creatinine concentrations using a method based on regression residuals (Mortamais et al. 2012) and applied previously in the EDEN cohort (Botton et al. 2016; Philippat et al. 2014). To limit the impact of outliers, phthalate metabolite concentrations were $\log_2$-transformed. The sum of DEHP metabolite concentrations (ΣDEHP) was calculated by summing molar concentrations of MEHP, MEHHP, MEOHP, and MECPP.

2.4. Placental tissue collection and DNA extraction

Placental tissue samples were obtained at delivery by the midwife or the technician of the study using a standardized procedure. Samples of around 5 mm^3 were collected from the fetal side, a few centimeters from the insertion of the cord, and immediately frozen at −80 °C. DNA was extracted using the QIAsymphony instrument (Qiagen, Germany) following the manufacturer's protocol. DNA concentration was determined by Nanodrop (ThermoFisher Scientific, USA) measurement and fluorescent quantification using PicoGreen (ThermoFisher Scientific, France). No sample was discarded due to low DNA concentration.

2.5. Placental DNA methylation assessment and quality control

As previously described (Abraham et al. 2018; Jedynak et al. 2021), whole-genome DNA methylation was measured at > 485,000 CpGs using the Infinium HumanMethylation450 BeadChips (Illumina, San Diego, CA, USA) following standard manufacturer's protocols. Raw intensities of fluorescent signals were processed with the Chip Analysis Methylation Pipeline (ChAMP) V2.14 (Morris et al., 2014; Tian et al., 2017). All samples but one passed initial quality control with an average of > 98% valid data points (detection p-value < 0.01). Filtering included removal of probes with detection p-values above 0.01 (52,692 probes), low numbers of measured events [beadcount < 3 in at least 5% of samples (44 probes)], probes not targeting a CpG (2,034 probes), probes associated with SNPs (50,829 probes, Zhou et al. 2017) or unspecific probes (9 probes, Nordlund et al. 2013). Methylation levels of individual CpGs were reported as continuous averaged β-value, representing the proportion of methylated alleles for each methylation site ranging from 0 (indicating that the site is completely unmethylated) to 1 (completely methylated) and were normalized in ChAMP using Beta MIxture Quantile (BMIQ) normalization (Teschendorff et al. 2013). To reduce the influence of outliers, methylation beta values above the 75th percentile + three interquartile ranges (IQRs) or below the 25th percentile − three IQRs were removed (in total, 0.39% of methylation values in our sample of 202 participants). 379,904 methylation sites remained after quality control, normalization, and filtering of outliers (Fig. 1). For each individual, pyrosequencing was used to measure methylation levels of four CpG sites of repetitive *Alu* elements (*Alu*) and long interspersed nucleotide elements 1 (LINE-1) (Yang 2004). One individual had no information on *Alu* methylation due to low signal-to-noise ratio and was removed from the analysis, yielding a final sample size of n = 201 for this study.

2.6. Placental cell heterogeneity estimation

We applied a recently proposed method to obtain reference-based estimates of placental cell composition (Yuan et al. 2021). In their work, Yuan et al. measured methylation profiles of six reference cell types [endothelial, Hofbauer, nucleated red blood cells (nRBC), stromal, syncytiotrophoblasts, and trophoblasts] in term placental tissues. Taking advantage of this cell-type specific reference provided in the R package *planet*, we applied the Robust Partial Correlations method implemented in the R package *EpiDISH* (Teschendorff et al. 2017) to the methylation data collected for a bigger EDEN sample (n = 668) that our study sample was derived from. This allowed us to obtain reference-based estimates of cell composition that were then used in our regression models as adjustment factors. If zero estimates were obtained for a cell type, they were considered as below the LOD and their values were imputed using the *impCoda* function from the R package *robCompositions* designed for compositional data and relying on an iterative regression-based procedure after KNN-initialization (Hron et al. 2010; Templ et al. 2011) (Supplementary Fig. 2).

2.7. Statistical analyses

2.7.1. Adjustment factors

We *a priori* selected factors that may affect both phthalate exposure and methylation marks in the placenta or the DNA methylation only. This included recruitment site (Nancy, Poitiers), maternal age (continuous), maternal pre-pregnancy BMI (< 18.5 kg/m^2, ≥ 18 - < 25 kg/m^2, ≥ 25 kg/m^2), maternal active smoking in the three months preceding pregnancy and during pregnancy (did not smoke, smoked before pregnancy, smoked before and during pregnancy, other), maternal education level ($<$ two years after high school, high school + two years, $\geq$ high school + three years), parity (nulliparous, $\geq$ one child), and season of conception (January-March, April-June, July-September, October-December). All analyses were additionally adjusted for technical factors related to the DNA methylation measurements [batch, plate, chip, and placental cell proportions for the epigenome-wide (EWAS) and global analysis of methylation profiles (GAMP); batch and plate for the study of repetitive *Alu* and LINE-1 elements]. Our analyses were not adjusted for gestational age at birth since it may be in the pathway between phthalate exposure and DNA methylation.

2.7.2. Associations with the global DNA methylation

We fitted one adjusted robust linear regression per phthalate biomarker to test the associations with global placental DNA methylation represented by the median methylation level of repetitive elements *Alu* or LINE-1. In the GAMP analysis we approximated the density and cumulative distribution functions of the methylation distribution using B-spline basis functions in order to characterize methylation profiles for each individual (Zhao et al. 2015a). Then we used the obtained B-spline coefficients as representatives of the individual overall methylation distribution. The variance component score test from the kernel machine framework served to test the adjusted associations between B-spline coefficients and concentrations of each phthalate biomarker. Using this method, we were able to evaluate whether pregnancy phthalate exposure changed the overall profile or distribution of DNA methylation for each individual instead of examining phthalates' effect on each CpG individually.

2.7.3. Associations with the CpG-specific DNA methylation

We performed an adjusted EWAS to assess associations between each phthalate metabolite concentration and DNA methylation at the level of individual CpG sites using robust linear regression (*MASS* R package Venables and Ripley 2002). p-values were calculated using Wald test from the *survey* R package (Lumley 2004) and corrected for false discovery rate (FDR) taking into account the number of CpGs tested for each chemical (Benjamini and Hochberg 1995). FDR corrected p-values
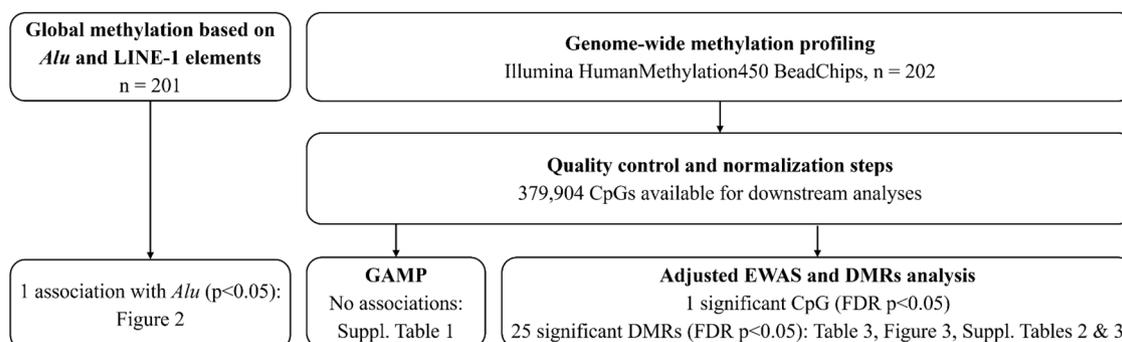

Fig. 1. Workflow of the statistical analyses performed in this study. Abbreviations: DMR = differentially methylated region. FDR = false discovery rate. GAMP = global analysis of methylation profiles.

below 0.05 were considered as statistically significant. Genomic inflation factor (λ) and Q-Q plots were generated using the *QCEWAS* R package (Van der Most et al. 2017) and the Bayesian inflation factor (BIF) was calculated using the *bacon* R/Bioconductor package (van Iterson et al., 2017). Gene annotations were based on Illumina's v1.2 annotation for the hg19 reference genome from the *IlluminaHumanMethylation450kanno.ilmn12.hg19* R/Bioconductor package (Hansen 2016) and information from the University of California, Santa Cruz (UCSC, https://genome.ucsc.edu) database.

2.7.4. Analysis of differentially methylated regions (DMRs)

To identify DMRs associated with phthalate biomarker concentrations we used the *comb-p* Python module (Pedersen et al. 2012). Using sliding windows, it combines p-values for CpGs detected in the EWAS accounting for their spatial correlations across the genome with a Stouffer-Liptak-Kechris correction (Kechris et al. 2010). Regional p-values are then adjusted for multiple testing by Šidák correction (Šidák 1967). DMRs with Šidák-corrected p-value below 0.05 and including at least two probes (p-value < 0.001 to initiate a region) at a maximum distance of 500 bp were considered significant. Basic information on genes encompassed by the DMRs identified as associated with the phthalate exposure were retrieved from the GeneCards Human Gene Database (Stelzer et al. 2016).

2.8. Research data and code

All analyses were conducted using R v. 4.0.5 (R Core Team and R Foundation for Statistical Computing 2020), RStudio v. 1.3.1106 (RStudio Team 2020) and Python v. 3.7.4 (van Rossum and Drake 2009). The data used in this study can only be provided upon reasonable request after approval by the EDEN steering committee. The code is available upon request to the corresponding authors. The statistical analysis plan for this study was pre-registered online (osf.io/2apqw).

3. Results

3.1. Study population characteristics and phthalate biomarker concentrations

Median maternal age was 29.1 years and median gestational duration was 40.0 weeks (Table 1). The minimal frequency of detection of phthalate metabolites was 98.5% and most of them were detected in 100% of the urine samples (Table 2). MEP was the most abundant biomarker (median standardized concentration: 117.8 μg/L) followed by MnBP (44.7 μg/L), MiBP (38.4 μg/L), and MECPP (38.1 μg/L). We observed strong correlations [Spearman's coefficient (rho) = 0.67] between the standardized concentrations of MCPP and MnBP and very strong correlations (0.84 ≤ rho ≤ 0.98) between the four individual DEHP metabolites (Supplementary Fig. 3).

3.2. Associations of phthalate biomarker urinary concentrations with global DNA methylation

Average methylation level was 16.2% (±1.1%) for *Alu* and 26.4% (±1.9%) for LINE-1. MBzP urinary concentrations were negatively associated with global methylation of the *Alu* element [β = −0.099, 95% confidence interval (95 %CI): −0.195; −0.002 for each doubling of MBzP concentration, Fig. 2]. None of the phthalate biomarker concentrations was significantly associated with the global methylation levels of the LINE-1 repetitive element nor with the overall profile or distribution of DNA methylation in the GAMP (Supplementary Table 1).

3.3. Associations between phthalate biomarker concentrations and individual CpG methylation levels

The p-value distributions of the CpGs included in the EWAS were

Table 1
Population characteristics for the 202 mother-son pairs included in the study and recruited between 2003 and 2006.

Characteristics	Distribution	
	n (%)	Median [25th, 75th centiles]
Center of recruitment		
Nancy	103 (51.0%)	
Poitiers	99 (49.0%)	
Season of conception		
January-March	44 (21.8%)	
April-June	41 (20.3%)	
July-September	57 (28.2%)	
October-December	60 (29.7%)	
Maternal active smoking in the 3 months preceding pregnancy and during pregnancy		
Did not smoke	127 (62.9%)	
Smoked before pregnancy	19 (9.4%)	
Smoked before and during pregnancy	26 (12.9%)	
Other[a]	30 (14.9%)	
Parity		
Nulliparous	88 (43.6%)	
≥ 1 child	114 (56.4%)	
Maternal level of education		
< 2 years after high school	93 (46.0%)	
high school + 2 years	43 (21.3%)	
≥ high school + 3 years	66 (32.7%)	
Maternal pre-pregnancy BMI[b]		
Underweight (< 18.5 kg/m^2)	19 (9.4%)	
Normal weight (≥ 18 - < 25 kg/m^2)	135 (66.8%)	
Overweight and obesity (≥ 25 kg/m^2)	48 (23.8%)	
Maternal age (years)		29.1 [25.6;33.0]
Gestational age at delivery (weeks)[c]		40.0 [38.9;41.0]

Abbreviations: BMI = body mass index. LMP = last menstrual period.
[a] Category "Other" referred to women that smoked at some point during pregnancy (during 1 or 2 out of 3 trimesters) but not during the whole pregnancy, or to women that smoked before pregnancy and at some point during pregnancy but not during the whole pregnancy (e.g., women who smoked in the 3 months preceding pregnancy and during the first trimester but quit smoking afterwards).
[b] Categorized according to the World Health Organization definitions.
[c] Based on the date of the LMP or gestational duration assessed by the obstetrician if it differed from the LMP-based estimate by more than 2 weeks.

close to the theoretical distribution as indicated by the BIF values (ranging from 0.98 to 1.05 depending on the phthalate), and were notably smaller compared to the genomic inflation factor values (0.90–1.33, Supplementary Fig. 4). We identified only one CpG (cg16039342) that was positively methylated (β = 0.02, 95 %CI: 0.01; 0.02) in association with an increase of MEHP concentration (FDR-corrected p-value < 0.05). The identified CpG mapped to the olfactomedin 2 (*OLFM2*) gene located on chromosome 19.

3.4. Regional DNA methylation analysis

The regional analysis identified 25 DMRs associated with phthalate biomarker concentrations during pregnancy (Šidák-corrected p-value < 0.05, Table 3). These 25 DMRs contained 131 CpGs and encompassed 22 protein coding genes, one RNA gene, and four intergenic regions (Fig. 3, Table 3, Supplementary Tables 2 and 3). For 22 (88%) detected DMRs, DNA methylation levels increased with increased phthalate biomarker concentrations.

Seven phthalate biomarkers were positively associated with

Table 2

Maternal urinary phthalate biomarker concentrations (n = 202).

Phthalate biomarker	Molecular weight	LOD (µg/L)	> LOD (%)	Measured concentrations Percentiles (µg/L)			Standardized concentrations[a] Percentiles (µg/L)		
				5th	50th	95th	5th	50th	95th
Monoethyl phthalate (MEP)	Low	0.6	100.0	23.1	121.5	1098.0	27.9	117.8	749.0
Mono-*iso*-butyl phthalate (MiBP)	Low	0.2	100.0	10.8	47.0	220.8	13.2	38.4	151.6
Mono-n-butyl phthalate (MnBP)	Low	0.2	100.0	11.5	60.0	625.5	14.2	44.7	689.2
Monobenzyl phthalate (MBzP)	High	0.3	100.0	3.4	19.6	102.9	4.9	17.7	89.8
Mono(3-carboxypropyl) phthalate (MCPP)	High	0.2	100.0	0.6	2.6	13.1	0.7	2.2	11.2
Mono(2-ethylhexyl) phthalate (MEHP)	High	0.5	98.5	0.9	8.6	42.6	1.3	7.2	30.1
Mono(2-ethyl-5-hydroxyhexyl) phthalate (MEHHP)	High	0.2	100.0	6.0	31.2	118.8	7.0	25.5	98.3
Mono(2-ethyl-5-oxohexyl) phthalate (MEOHP)	High	0.2	100.0	5.0	24.6	83.0	5.7	20.2	76.8
Mono(2-ethyl-5-carboxypentyl) phthalate (MECPP)	High	0.2	100.0	11.7	43.1	176.9	12.2	38.1	142.6
Molar sum of DEHP metabolites (ΣDEHP)[b]	High	NA	NA	0.1	0.4	1.4	0.1	0.3	1.1
Monocarboxy-*iso*-octyl phthalate (MCOP)	High	0.2	99.5	1.0	3.6	19.4	1.1	3.8	18.8
Monocarboxy-*iso*-nonyl phthalate (MCNP)	High	0.2	99.5	0.5	1.7	16.8	0.6	1.4	10.5

Phthalate metabolite concentrations are displayed in µg/L for all compounds except for ΣDEHP for which the concentrations are presented as µmol/L.

Abbreviations: LOD = limit of detection. NA = not applicable.

[a] Measured concentrations were standardized on sampling conditions (hour of sampling, day of sampling, year of sample analysis at the CDC, gestational age at collection, duration of storage at room temperature before freezing) and creatinine concentration using a method based on regression residuals (Mortamais et al. 2012; Philippat et al. 2014).

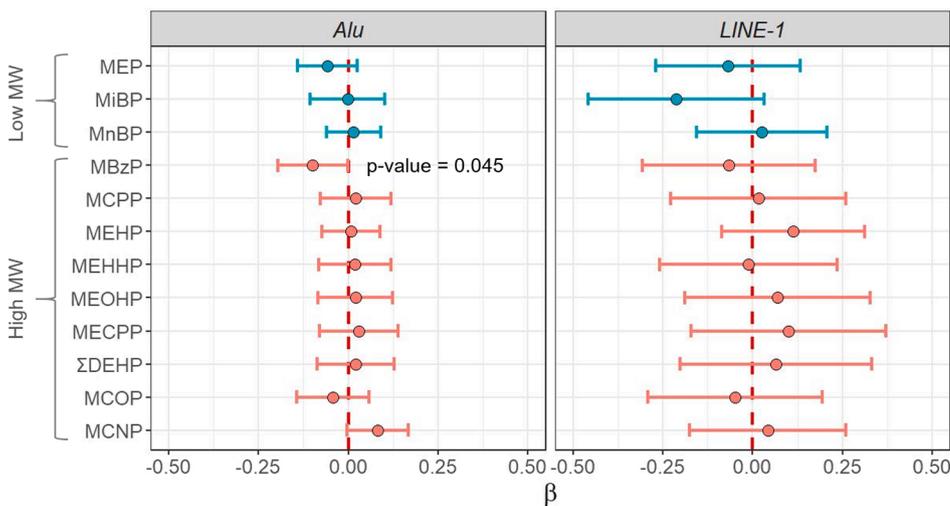[b] ΣDEHP was calculated by summing molar concentrations of MEHP, MEHHP, MEOHP, and MECPP.

Fig. 2. Adjusted associations between urinary concentrations of phthalates biomarkers and methylation levels of the repetitive elements *Alu* and LINE-1 during pregnancy (n = 201). Circles represent β regression coefficient estimates reported with 95% CIs and correspond to a change in the global DNA methylation level for doubling of the urinary biomarker concentration. Colors represent biomarkers of low (blue) or high (red) molecular weight phthalates. Regression models were adjusted for recruitment center, maternal active smoking in the three months preceding pregnancy and during pregnancy, maternal age, parity, maternal education level, maternal pre-pregnancy BMI, season of conception, batch and plate. Abbreviations: BMI = body mass index. CI = confidence interval. MBzP = monobenzyl phthalate. MCNP = monocarboxy-*iso*-nonyl phthalate. MCOP = monocarboxy-*iso*-octyl phthalate. MCPP = mono(3-carboxypropyl) phthalate. MECPP = mono(2-ethyl-5-carboxypentyl) phthalate. MEHHP = mono(2-ethyl-5-hydroxyhexyl) phthalate. MEHP = mono(2-ethylhexyl) phthalate. MEOHP = mono(2-ethyl-5-oxohexyl) phthalate. MEP = monoethyl phthalate. MiBP = mono-*iso*-butyl phthalate. MnBP = mono-n-butyl phthalate. MW = molecular weight. ΣDEHP = molar sum of di(2-ethylhexyl) phthalate metabolites (MEHP, MEHHP, MEOHP, MECPP).

placental DNA methylation while two (MBzP and MEHP) showed both positive and negative associations with DNA methylation. MEP was positively associated with two DMRs mapping to two genes: grancalcin (*GCA*) and farnesyl-diphosphate farnesyltransferase 1 (*FDFT1*). MiBP concentrations were associated with increased methylation of the replication initiator 1 (*REPIN1*). MnBP concentrations were positively associated with six DMRs that encompassed four protein coding genes including forkhead box S1 (*FOXS1*), heparan sulfate-glucosamine 3-sulfotransferase 3B1 (*HS3ST3B1*), spi-B transcription factor (*SPIB*), solute carrier family 17 member 9 (*SLC17A9*), an RNA gene (*MGC12916*) coding uncharacterized protein MGC12916, and two intergenic regions.

DEHP metabolites were associated with eight DMRs. MEHP metabolite was positively associated with epidermal growth factor receptor pathway substrate 8 like 1 (*EPS8L1*), myosin heavy chain 3 (*MYH3*), and *OLFM2*. It was negatively associated with zinc finger and SCAN domain containing 16 (*ZSCAN16*) gene methylation. MEOHP concentrations were linked with increased DNA methylation of two genes: transcription factor 21 (*TCF21*), TIAM Rac1 associated GEF 2 (*TIAM2*), and one

intergenic region, while MECPP concentrations were positively associated with only one DMR that encompassed the fibroblast growth factor 12 (*FGF12*).

Pregnancy concentrations of MCNP were positively associated with the largest identified DMR (22 probes) and encompassed the heat shock protein (HSP) family A member 1A (*HSPA1A*) and *HSPA1* like (*HSPA1L*) genes. This phthalate biomarker was also positively associated with carbonic anhydrase 5B (*CA5B*) and poly(ADP-ribose) polymerase family member 10 (*PARP10*). MCPP concentrations were positively associated with the methylation of ADP-ribosyltransferase 5 (*ART5*) gene and of one intergenic region. Finally, MBzP was positively associated with methylation of rap guanine nucleotide exchange factor like 1 (*RAP-GEFL1*) and negatively associated with N-acetylated alpha-linked acidic dipeptidase 2 (*NAALAD2*) and SRY-box transcription factor 21 (*SOX21*).

4. Discussion

Herein we explored the epigenome-wide associations between

Table 3

DMRs associated with pregnancy concentrations of phthalate metabolites (25 DMRs, Šidák-corrected p-value < 0.05, n = 202, 379,904 CpGs).

Phthalate metabolite	Molecular weight	Gene[a]	DMR (chromosome:start-end)	No. of CpGs	SLK p-value	Šidák p-value	Direction of association
Monoethyl phthalate (MEP)	Low	FDFT1	chr8:11659961-11660110	6	4.02E−09	1.03E−05	+
Monoethyl phthalate (MEP)	Low	GCA	chr2:163200476-163200638	6	4.27E−10	1.00E−06	+
Mono-*iso*-butyl phthalate (MiBP)	Low	REPIN1	chr7:150065170-150065256	6	1.70E−12	7.51E−09	+
Mono-n-butyl phthalate (MnBP)	Low	FOXS1	chr20:30433405-30433513	2	4.37E−07	1.53E−03	+
Mono-n-butyl phthalate (MnBP)	Low	HS3ST3B1/ MGC12916	chr17:14206871-14207037	6	2.91E−08	6.67E−05	+
Mono-n-butyl phthalate (MnBP)	Low	SLC17A9	chr20:61583979-61584073	6	1.85E−09	3.53E−06	+
Mono-n-butyl phthalate (MnBP)	Low	SPIB	chr19:50931515-50931623	3	1.17E−08	4.12E−05	+
Mono-n-butyl phthalate (MnBP)	Low		chr12:115135238-115135369	5	3.95E−12	7.50E−09	+
Mono-n-butyl phthalate (MnBP)	Low		chr3:185000648-185000761	3	3.75E−08	1.26E−04	+
Mono(2-ethylhexyl) phthalate (MEHP)	High	EPS8L1	chr19:55598874-55599030	2	2.30E−08	5.59E−05	+
Mono(2-ethylhexyl) phthalate (MEHP)	High	MYH3	chr17:10541530-10541621	2	6.81E−07	2.84E−03	+
Mono(2-ethylhexyl) phthalate (MEHP)	High	OLFM2	chr19:9965044-9965172	2	6.95E−09	2.06E−05	+
Mono(2-ethylhexyl) phthalate (MEHP)	High	ZSCAN16	chr6:28092239-28092421	7	3.35E−09	6.99E−06	−
Mono(2-ethyl-5-carboxypentyl) phthalate (MECPP)	High	FGF12	chr3:192445514-192445539	3	1.07E−06	1.62E−02	+
Mono(2-ethyl-5-oxohexyl) phthalate (MEOHP)	High	TCF21	chr6:134210138-134210308	7	4.75E−09	1.06E−05	+
Mono(2-ethyl-5-oxohexyl) phthalate (MEOHP)	High	TIAM2	chr6:155537901-155538056	5	3.50E−08	8.58E−05	+
Mono(2-ethyl-5-oxohexyl) phthalate (MEOHP)	High		chr6:32828996-32829145	4	1.70E−08	4.34E−05	+
Monobenzyl phthalate (MBzP)	High	NAALAD2	chr11:89867653-89867820	5	1.74E−07	3.96E−04	−
Monobenzyl phthalate (MBzP)	High	RAPGEFL1	chr17:38347603-38347817	4	2.97E−10	5.26E−07	+
Monobenzyl phthalate (MBzP)	High	SOX21	chr13:95364510-95364676	6	1.17E−08	2.68E−05	−
Mono(3-carboxypropyl) phthalate (MCPP)	High	ART5	chr11:3663491-3663843	8	1.47E−12	1.80E−09	+
Mono(3-carboxypropyl) phthalate (MCPP)	High		chr12:115135333-115135369	5	5.33E−11	1.01E−07	+
Monocarboxy-*iso*-nonyl phthalate (MCNP)	High	CA5B	chrX:15756372-15756408	4	4.65E−08	4.90E−04	+
Monocarboxy-*iso*-nonyl phthalate (MCNP)	High	HSPA1A/HSPA1L	chr6:31782873-31783546	22	9.83E−18	5.55E−15	+
Monocarboxy-*iso*-nonyl phthalate (MCNP)	High	PARP10	chr8:145061291-145061319	2	9.85E−09	1.34E−04	+

EWAS regression models, on which the DMR analysis was based, were adjusted for recruitment center, maternal active smoking in the three months preceding pregnancy and during pregnancy, maternal age, parity, maternal education level, maternal pre-pregnancy BMI, season of conception, batch, plate, chip, and estimated placental cell-type proportions.

Abbreviations: BMI = body mass index. DMR = differentially methylated region. SLK = Stouffer-Liptak-Kechris correction.

[a] University of California, Santa Cruz Genome Browser (https://genome.ucsc.edu).

concentrations of phthalate exposure biomarkers during pregnancy and placental DNA methylation. As for global DNA methylation, increased maternal concentrations of MBzP were associated with decreased methylation of the repetitive *Alu* element. With regard to other analyses, in the following discussion we will focus on the DMR results as biological functions are associated rather with genomic regions than with single CpGs (Svendsen et al. 2016). All studied phthalate biomarkers except for MCOP and MEHHP were associated with at least one DMR. For most of the identified regions (n = 22, 88%), DNA methylation levels were increased. Identified DMRs encompassed 23 genes encoding heat shock proteins, transcription factors, and nucleotide exchange factors, among others.

4.1. Low molecular weight (LMW) phthalates

LMW phthalates are frequently found in cosmetics and personal care products such as shampoos, perfumes, aftershaves, or lotions (Dodson et al. 2012). Pregnancy concentrations of LMW phthalate metabolites, including MEP, MnBP, and MiBP [either individually or the molar sum of the three metabolites (i.e., ∑LMW)], have been previously associated with differential DNA methylation in placenta in a study considering two candidate imprinted genes: *H19* and *IGF2* (n = 179, LaRocca et al. 2014). The authors reported DNA methylation loss within the *IFG2* DMR0 associated with increased pregnancy concentrations of MEP and within the *H19* and *IFG2* DMR0 associated with increased ∑LMW phthalate biomarker concentrations. There were only eight common probes between LaRocca et al. and our study (one for *IFG2* DMR0 and seven for *IFG2* DMR2) and we did not identify any of them to be
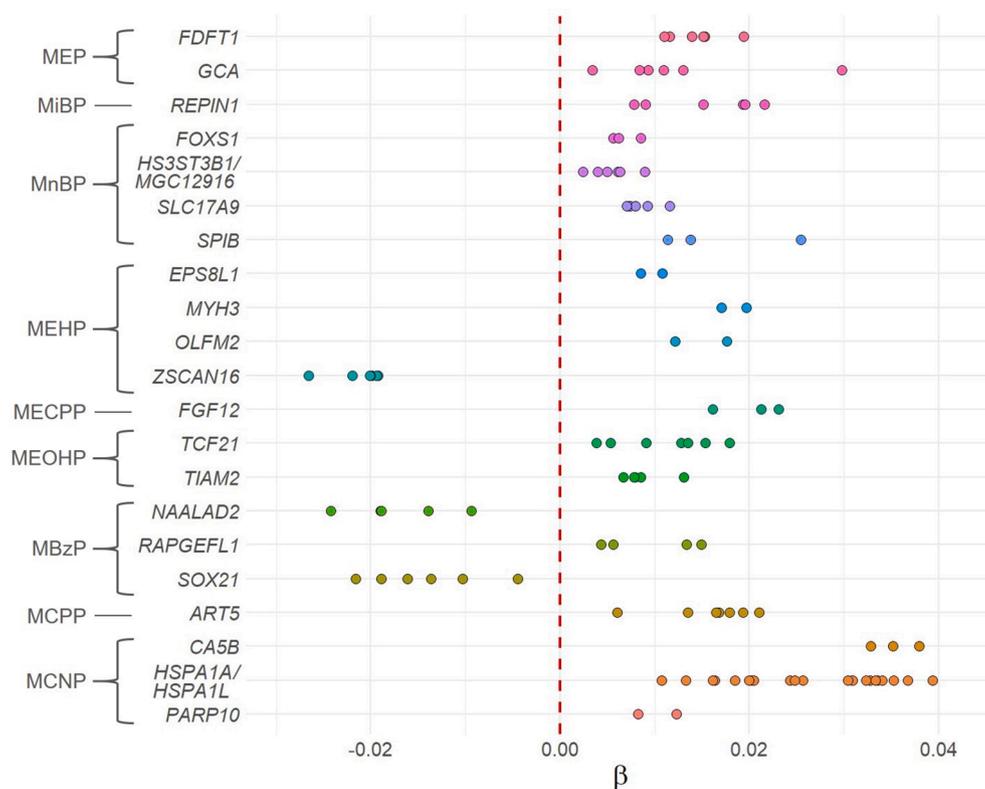
Fig. 3. β regression coefficient estimates according to genes encompassed by DMRs identified as associated with phthalate biomarkers (Šidák-corrected p-value < 0.05, n = 202, 379,904 CpGs). Colors represent genes. Circles represent CpGs mapping to genes within identified DMRs. β coefficient estimates correspond to a change in the DNA methylation level for doubling of the urinary exposure concentration. EWAS regression models, on which the DMR analysis was based, were adjusted for recruitment center, maternal active smoking in the three months preceding pregnancy and during pregnancy, maternal age, parity, maternal education level, maternal pre-pregnancy BMI, season of conception, batch, plate, chip, and estimated placental cell-type proportions. Abbreviations: BMI = body mass index. DMR = differentially methylated region. MBzP = monobenzyl phthalate. MCNP = mono-carboxy-*iso*-nonyl phthalate. MCPP = mono (3-carboxypropyl) phthalate. MECPP = mono(2-ethyl-5-carboxypentyl) phthalate. MEOHP = mono(2-ethyl-5-oxohexyl) phthalate. MEP = monoethyl phthalate. MiBP = mono-*iso*-butyl phthalate. MnBP = mono-n-butyl phthalate. MEHP = mono(2-ethyl-hexyl) phthalate.

associated with MEP (p-values not corrected for FDR ≥ 0.17). However, in our study 91 additional CpGs mapped to the *IGF2* gene and for four of them increased MEP concentrations were associated with DNA methylation change (p-value not corrected for FDR < 0.05, data not shown); however, in contrast to LaRocca et al., these associations were positive. Result discrepancies between the two studies may come from different methodologies used for DNA methylation assessment (pyrosequencing vs. BeadChip technology), distinct populations, timing of urine collection (< 16 gestational weeks in LaRocca et al. compared to 22–29 weeks in our study), and different phthalate biomarker levels (geometric mean of MEP concentration equaled 76.2 µg/L in LaRocca et al. compared to 124.1 µg/L in the present study). The set of adjustment factors also differed between the two studies (LaRocca et al. adjusted only for child sex, maternal smoking, and maternal age).

As for tissues other than placenta, a study performed in the CHA-MACOS cohort reported negative associations between MiBP and MnBP assessed in early pregnancy (13 weeks) and methylation of cord blood LINE-1, and between MEP and ∑LMW assessed in late pregnancy (26 weeks) and *Alu* methylation (n = 239, Huen et al. 2016). In our study we did not observe any association between LMW phthalates and methylation of the repetitive elements and the discrepancy may be related to different biological matrices used (cord blood vs. placenta) or different exposure assessment windows (13 or 26 gestational weeks in Huen et al. compared to 22–29 weeks in our study).

Although not affecting global DNA methylation, we found MEP concentrations being positively associated with two DMRs encompassing *FDFT1* and *GCA* genes. *FDFT1* encodes an enzyme important in cholesterol biosynthesis (O'Leary et al., 2016) and *GCA* encodes a calcium-binding protein abundant in neutrophils and macrophages and plays a role in the innate immune response (Stelzer et al. 2016). Both *FDFT1* and *GCA* genes have been previously associated with maternal smoking. Placental DNA methylation of *FDFT1* has been shown to be associated with pregnancy tobacco use (direction and magnitude of the association depended on the smoking status) in a previous study on the EDEN cohort (n = 668, Rousseaux et al. 2020). Moreover, *FDFT1* has

been found to be up-regulated in lung tissue obtained from smokers with lung adenocarcinoma (Pintarelli et al. 2019). As for *GCA*, one study with a relatively small sample size showed an increase of its placental methylation associated with maternal smoking during pregnancy (n = 36, Suter et al. 2011), however an aforementioned study with a bigger sample size carried on the EDEN cohort did not replicate this result (n = 668, Rousseaux et al. 2020). Since cigarette smoke may contain diethyl phthalate (DEP) (Moldoveanu and St. Charles, 2007), the parent compound of MEP, this could partially explain the observed association between this phthalate metabolite and differential methylation of *GCA* and *FDFT1* previously linked with maternal smoking. Our results could also suggest that these genomic locations are particularly sensitive to environmental exposure.

In our population, another LMW phthalate, MnBP, was positively associated with six DMRs. Maternal MnBP urinary concentrations have been previously associated with higher expression of inflammation-related genes in placenta in a study relying on a candidate gene approach (n = 2469, Wang et al. 2020b). In our study, we have not detected differential methylation of any of inflammation-related genes, which may be explained by the fact that gene expression does not necessarily correlate with DNA methylation levels. Instead, we identified an increase of DNA methylation of four protein coding genes (*FOXS1*, *HS3ST3B1*, *SLC17A9*, *SPIB*), one long non-coding RNA gene (*MGC12916*), and two intergenic regions. *FOXS1* and *SPIB* encode transcription factors, HS3ST3B1 protein plays a role in nucleotide binding, and SLC17A9 is a transmembrane protein involved in the transport of small molecules.

Lastly, we found maternal MiBP concentrations being associated with increased placental DNA methylation of *REPIN1* gene encoding protein facilitating DNA binding. To the best of our knowledge, none of the genes identified in our study as associated with pregnancy MnBP or MiBP concentrations has been previously described in the context of epigenetic modifications or functioning of the placenta.

4.2. High molecular weight (HMW) phthalates

HMW phthalates are used as plasticizers in products such as food packaging, plastic bags, vinyl plastics used in flooring, toys, and intravenous tubing (Hauser and Calafat 2005).

4.2.1. DEHP metabolites

DEHP and its metabolite MEHP may alter placental homeostasis by disrupting trophoblast differentiation, invasion, oxidative stress response, immuno-modulation, and endocrine function (reviewed by Martínez-Razo et al. 2021). Regarding potential effect of DEHP metabolites on the placental DNA methylation, LaRocca et al. showed MEOHP and ΣDEHP to be negatively associated with *IGF2* DMR0 methylation (LaRocca et al. 2014). After stratification for sex, they also reported negative associations between MECPP, MEHHP, and MEHP and *IGF2* DMR0 methylation in females. In our study restricted to males, after analysis of all CpGs mapping to the *IGF2* gene (99 CpGs), we found a few positive associations with MECPP (3 CpGs), MEHHP (2), MEHP (3), MEOHP (2), and ΣDEHP (3) and one negative association with MEHP (p-values not corrected for FDR < 0.05, data not shown). However, none of these CpGs was located within the *IGF2* DMR0.

Two other studies investigated associations between DEHP metabolites and placental genes methylation. One study using placentas from fetal growth restricted and normal growth newborns focused on candidate imprinted genes *IGF2* and *AHRR* (n = 181, Zhao et al. 2016). The authors found maternal urinary concentrations of MEHHP, MEOHP, and the molar sum of MEHHP, MEHP, and MEOHP to be negatively associated with *IGF2* methylation. The second study on fetal growth restricted infants reported decrease in methylation of placental LINE-1 element in association with urinary phthalate DEHP metabolites concentrations (MEHHP and molar sum of MEHHP, MEHP, and MEOHP) (n = 65, Zhao et al. 2015b). In our study, we have not identified any of these genes to be associated with DEHP metabolites, which may be explained by the fact that associations observed by Zhao et al. were present only in the growth restricted and not in normal growth newborns.

As for tissues other than placenta, a few studies investigated associations between maternal urinary concentrations of DEHP metabolites and DNA methylation in cord blood. A study carried on the CHAMACOS cohort and focusing on ten candidate imprinted genes reported positive associations between MECPP, MEHHP, MEOHP, and ΣDEHP and DNA methylation averaged across seven CpGs of maternally expressed 3 (*MEG3*) gene (n = 296, Tindula et al. 2018). No effect was reported for the *MEG3* expression. Inversely, a study focusing on candidate genes playing a role in metabolism, growth, or development showed ΣDEHP to be associated with decreased cord blood methylation of peroxisome proliferator-activated receptor alpha (*PPARA*), a gene encoding nuclear receptor that regulates fatty acid metabolism. In our study we did not observe differential methylation of either *MEG3* or *PPARA*, which may be explained by different biological matrices (cord blood vs. placenta), distinct methodologies of DNA methylation assessment (pyrosequencing vs. BeadChip) or distinct populations and adjustment factors or timing of exposure assessment (8–14 gestational weeks in Montrose et al. and averaged concentrations from 13th and 26th weeks in Tindula et al. compared to 22–29 weeks in our study). Lastly, an epigenome-wide study on the CHAMACOS cohort identified 27 cord blood DMRs predominantly positively associated with phthalate biomarkers (n = 336, Solomon et al. 2017). Identified DMRs were primarily associated with individual and summary measurements of DEHP metabolites assessed in the 26th week of gestation. They encompassed several genes related to hormonal balance, male fertility, metabolic health and cancer, however none of them was common with our study. Instead, we found all DEHP biomarkers except for MEHHP and ΣDEHP to be associated with at least one DMR. MEHP was positively associated with three genes: *EPS8L1*, *MYH3*, and *OLFM2*. *OLFM2* is involved in smooth muscle differentiation and was the only gene also detected in the EWAS. *MYH3* encodes one of the myosin heavy chains playing a role in motor activity while the exact

function of the protein encoded by the *EPS8L1* gene is unknown. In contrast, MEHP was associated with decreased DNA methylation of the *ZSCAN16* gene coding DNA-binding transcription factor. *ZSCAN16* methylation has been shown to be increased in association with high total phthalate exposure (sum of concentrations of 23 metabolites including MEHP) in first trimester placentas of women undergoing elective terminations (among 244 other genes none of which was identified in our study) (Grindler et al. 2018). However, the latter result should be interpreted with caution since Grindler et al. relied on 16 placentas only while analyzing as many as 834,015 CpG sites; they also did not formally correct for multiple comparisons and used a relatively low threshold to detect significant DMRs (p-value < 0.005). Moreover, the authors did not adjust their analyses for potential confounders. *ZSCAN16* differential DNA methylation has been also associated with maternal smoking in a study conducted in the EDEN cohort (n = 668, Rousseaux et al. 2020). However, in contrast to our results, this association was positive. To the best of our knowledge, the presence of DEHP or any of its metabolites in tobacco smoke has not been reported. Nevertheless, it has been demonstrated that DEHP exposure may induce production of reactive oxygen species (ROS) in both *in vitro* models as well as in pregnant women with high levels of DEHP metabolites detected in urine (reviewed by Martínez-Razo et al. 2021). Moreover, high pregnancy urinary concentrations of DEHP metabolites have been linked with placental overexpression of metallothioneins, proteins that show cellular antioxidative properties (Li et al. 2016). Because tobacco smoke is also a well-known factor related to ROS production and was recently linked to oxidative damages in the placenta (reviewed by Suter and Aagaard 2020), the common effect of DEHP metabolites and smoking on the *ZSCAN16* methylation may be explained by the production of ROS linked to these two factors. Our results may also suggests particular sensitivity of this locus to environmental exposures.

In the present study, concentrations of another DEHP metabolite, MEOHP, were associated with increased DNA methylation of two genes (*TCF21* and *TIAM2*) and one intergenic region. The protein encoded by *TIAM2* gene is a nucleotide exchange factor suspected to play a role in neural cell development (O'Leary et al., 2016). The impact of epigenetic modifications of this gene in placenta or cord blood have not been studied so far. *TCF21* is a transcription factor and a tumor suppressor. Increased placental methylation of this gene was associated with maternal smoking in the EDEN cohort (n = 668, Rousseaux et al. 2020). Again, the common mechanism linking maternal MEOHP concentrations and smoking status with increased placental DNA methylation of the *TCF21* gene may be related to the production of ROS. Finally, MEOHP have been previously shown to be negatively associated with the expression of the inflammation-related genes (*CRP*, *MCP-1*, *CD68*) in female placentas in a study relying on a candidate approach (n = 2469, Wang et al. 2020b). In our study restricted to boys we did not identify such associations.

The last association we detected for DEHP metabolites was positive and involved pregnancy MECPP concentrations and DNA methylation of *FGF12*, a gene from the FGF family involved in a variety of biological processes including embryonic development, cell growth, morphogenesis, tissue repair, and tumor growth and invasion (O'Leary et al., 2016). The specific function of the *FGF12* gene has not yet been determined nor has it been previously described in the context of epigenetic modifications or functioning of the placenta.

4.2.2. HMW phthalates other than DEHP metabolites

To date, none of the studies on placenta reported differential DNA methylation associated with maternal concentrations of HMW phthalates other than DEHP metabolites (LaRocca et al. 2014; Zhao et al. 2015b, 2016). As for cord blood, a study by Montrose et al. showed MBzP and MCPP being negatively associated with methylation of *PPARA* (MBzP) and *IGF2*, *PPARA*, and LINE-1 (MCPP). However, after stratification for sex, some of these associations appeared to be female-specific or were not retained (Montrose et al. 2018), what may partially explain

why we did not observe such associations in our study restricted to boys. Two other studies focused on the cord blood DNA methylation of LINE-1 and *Alu* repetitive elements. They reported negative associations between MBzP concentrations assessed in early pregnancy and LINE-1 methylation (Huen et al. 2016) as well as between MnBP concentrations and *Alu* methylation in males and MBzP and *Alu* methylation in females (n = 106, Huang et al. 2018). The latter association was also detected in our study restricted to boys which may suggest a potential mechanism through which exposure to benzylbutyl phthalate, the parent compound of MBzP, may affect placental health.

As for other studies involving MBzP, it has been previously shown to alter placental expression of inflammation-related genes (*TNF-α*, *MCP-1*, and *CD68*) in males (n = 2469, Wang et al. 2020b), but none of these genes was differentially methylated in our study. Instead, we identified an increase of the DNA methylation of the *RAPGEFL1* gene and a decrease of methylation of two other genes (*NAALAD2* and *SOX21*) associated with this phthalate. *RAPGEFL1* is a nucleotide exchange factor showing signal transducer activity and *NAALAD2* has neuropeptide cleaving function (Stelzer et al. 2016), but with no proven role in placental functioning. The third gene associated with MBzP was the transcription factor *SOX21* that encodes a protein regulating placentation and differentiation of the trophoblast (Mrema et al. 2013; Ullah et al. 2020), among other tissues. Placental *SOX21* was found to be downregulated in association with total phthalates (sum of concentrations of 23 metabolites including MBzP) in the above cited study on women undergoing elective pregnancy terminations during first trimester (n = 16, Grindler et al. 2018). Therein, decreased *SOX21* expression was accompanied by an increase of DNA methylation of one CpG mapping to this gene, while in our study we observed DNA methylation loss within this gene.

Another DMR identified in our study was positively associated with MCNP concentrations and was the largest (22 CpGs) among those detected. This DMR encompasses *HSPA1A* and *HSPA1L* genes encoding heat shock proteins responsible for various physiological processes (e.g., protein refolding and degradation) and involved in response to cellular stress. Increased DNA methylation of *HSPA1A/ HSPA1L* genes has been reported in intrauterine growth restriction placentas, compared to their normal twin counterparts (n = 8, Roifman et al. 2016). Additionally, increased placental *HSPA1A* mRNA and protein levels have been reported in preeclamptic pregnancies (n = 8, Wang et al. 2020a) and in placental vascular disease (n = 62, Liu et al. 2008). Of interest, MCNP has been previously associated with the placenta weight to birth weight ratio in another study from the EDEN cohort relying on 457 mother-son pairs (Philippat et al. 2019), which may suggest an effect of this phthalate on fetal growth and development, potentially mediated by the epigenetic modifications in placenta. We observed two additional positive associations for MCNP with *CA5B* and *PARP10* genes encoding proteins showing carbonate dehydratase activity and responsible for gene transcription regulation, respectively (O'Leary et al., 2016). To the best of our knowledge, none of these genes has been studied in the context of placental function.

Lastly, we detected a positive association between MCPP concentration and DNA methylation of one intergenic region and the *ART5* gene encoding protein responsible for protein function regulation. Again, there are no studies linking epigenetic modifications of *ART5* to placental outcomes.

5. Strengths and limitations

The present study is the first genome-wide analysis of differentially methylated probes and regions in placenta collected at birth in relation to pregnancy exposure to phthalates. In our analyses, we corrected for the number of CpGs tested for each chemical but we did not account for the number of tested phthalate metabolites, which might have led to identification of false positive associations. It should be noted though that for 11 studied phthalate metabolites, all except for MEP showed

moderate (MCNP, MCOP, MBzP, MiBP), strong (MCPP, MnBP) or very strong (DEHP metabolites) correlation with at least one other phthalate metabolite. When this was taken into account, the effective number of independent exposures (formula adapted from Li et al. 2012) dropped from 11 to six (data not shown). Taken together with the lack of other epigenome-wide studies on the effects of phthalates on placental DNA methylation, cautious interpretation of the results is required and replication studies are needed to confirm our findings. Moreover, the fact that the observed changes in DNA methylation cannot be directly translated to gene expression (Lim et al. 2017) may impede the interpretation of how our results may link to pregnancy complications or health outcomes later in life.

In this study, we followed two statistical approaches providing complementary information. The EWAS produces CpG-specific effect estimates that can be directly compared with the estimates observed in other studies and that can be used in *meta*-analyses. On the other hand, DMR analysis takes into account the location of the differentially methylated CpGs in the genome and their potential interdependence. Although the EDEN mother–child cohort is well established and provides information on a broad range of potential confounders, residual confounding by factors not considered in our analysis (e.g., genetics, ancestry, or maternal behaviors such as diet or physical activity during pregnancy) cannot be excluded. Phthalate exposure was assessed only for boys which, while not being a source of bias, limits the generalizability of our conclusions for female offspring. This is especially important because previous studies have reported sex-specific effects of some phthalates on methylation and expression of genes in placenta and cord blood (Huang et al. 2018; LaRocca et al. 2014; Montrose et al. 2018). Finally, phthalate biomarkers were assessed in a single spot urine sample. Given the short half-life of phthalates (Casas et al. 2018) and temporal variability in behaviors linked to exposure (e.g., food intake, use of personal care products), we cannot rule out exposure misclassification, attenuation bias, and power reduction (Perrier et al. 2016). Exposure misclassification depends on the temporal variability of urinary concentrations which, as shown in Supplementary Table 4, varies across phthalate metabolites. The highest attenuation bias is expected for the metabolites with the highest temporal variability (i.e., with the lowest intraclass correlation coefficient) which, for our study, include the metabolites with the highest molecular weight (MCNP, MCOP, MCPP, MECPP, MEHHP, MEHP, and MEOHP) (Supplementary Table 4).

6. Conclusions

Herein we explored the epigenome-wide effects of pregnancy exposure to selected phthalates on DNA methylation in placentas collected at birth. We found that MBzP concentrations may be associated with decreased methylation of repetitive *Alu* elements. Moreover, most of the studied phthalates were associated with increased DNA methylation of several DMRs. For two metabolites (MBzP and MEHP), decreased methylation of three DMRs was also observed. Identified regions encompassed 23 genes encoding heat shock proteins, transcription factors, and nucleotide exchange factors, among others. Of interest, four genes have been previously identified as associated with maternal smoking, suggesting that these genomic regions might be particularly sensitive to the effect of environmental contaminants. Presented results suggest epigenetic mechanisms by which pregnancy exposure to phthalates could affect fetal development; however, additional studies are needed to confirm our results.

CRediT authorship contribution statement

Paulina Jedynak: Formal analysis, Investigation, Writing – original draft, Writing – review & editing. **Jörg Tost:** Data curation, Formal analysis, Resources. **Antonia M. Calafat:** Formal analysis, Resources. **Ekaterina Bourova-Flin:** Methodology. **Lucile Broséus:** Methodology. **Florence Busato:** Data curation, Formal analysis, Resources. **Anne**

Forhan: Data curation. **Barbara Heude:** Conceptualization, Data curation, Resources. **Milan Jakobi:** Data curation, Software. **Joel Schwartz:** Formal analysis, Methodology. **Rémy Slama:** Funding acquisition. **Daniel Vaiman:** Formal analysis, Resources, Methodology. **Johanna Lepeule:** Conceptualization, Data curation, Formal analysis, Resources, Funding acquisition, Investigation, Methodology, Software, Supervision, Writing – review & editing **Claire Philippat:** Conceptualization, Formal analysis, Funding acquisition, Investigation, Supervision, Writing – review & editing.

Declaration of Competing Interest

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

Appendix A. Supplementary material

Supplementary data to this article can be found online at https://doi.org/10.1016/j.envint.2021.107054.

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