

Exposure of pregnant women to organophosphate insecticides and child motor inhibition at the age of 10-12 years evaluated by fMRI

Anne-Claire Binter, Elise Bannier, Dave Saint-Amour, Grégory Simon, Christian Barillot, Christine Monfort, Sylvaine Cordier, Fabienne Pelé, Cécile Chevrier

► To cite this version:

Anne-Claire Binter, Elise Bannier, Dave Saint-Amour, Grégory Simon, Christian Barillot, et al.. Exposure of pregnant women to organophosphate insecticides and child motor inhibition at the age of 10–12 years evaluated by fMRI. Environmental Research, 2020, 188, pp.109859. 10.1016/j.envres.2020.109859. inserm-02932111

HAL Id: inserm-02932111 https://inserm.hal.science/inserm-02932111

Submitted on 10 Sep 2020

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

- 1 Exposure of pregnant women to organophosphate insecticides and child motor
- 2 inhibition at the age of 10 to 12 years evaluated by fMRI
- Binter A.C¹, Bannier E², Saint-Amour D³, Simon G⁴, Barillot C⁵, Monfort C¹, Cordier S¹, Pelé F^{1,6*},
 Chevrier C^{1*}
- 5 ¹ Univ Rennes, Inserm, EHESP, Irset (Institut de recherche en santé, environnement et travail) -
- 6 UMR_S 1085, F-35000 Rennes, France
- 7 ² Univ Rennes, CHU Rennes, CNRS, Inria, Inserm, IRISA UMR 6074, Empenn ERL U 1228, F-35000
- 8 Rennes, France
- 9³ Department of Psychology, Université du Québec à Montréal, Montréal, Canada
- 10 ⁴ ISTS EA 7466, University of Caen Normandie, Caen, France
- ⁵ Univ Rennes, CNRS, Inria, Inserm, IRISA UMR 6074, Empenn ERL U 1228, F-35000 Rennes, France
- 12 ⁶ Univ Rennes, Inserm, CIC 1414, Rennes, France
- 13 * Equal contributors
- 14 Correspondence address: Inserm UMR 1085 Irset, 9 Avenue du Pr. Léon Bernard, 35000 Rennes,
- 15 France, <u>anne-claire.binter@gmail.com</u>
- 16 <u>Conflict of interest:</u> The authors declare they have no actual or potential competing financial
- 17 interests.

18 Highlights

19 • Organophosphates (OP) are widely used in agricultural and domestic settings.

• Adverse neuro-outcomes were linked to OP exposure but underpinnings are unknown.

• Functional MRI was performed for 95 children aged 10-12 years.

- Activity in frontal regions decreased with increasing levels of OP metabolites.
- OP may be associated with altered activity in brain regions related to inhibition.

24 <u>Abstract</u>

Background: Organophosphate pesticides (OP) are widely used for both agricultural and domestic
purposes. Epidemiological studies suggest neurotoxicity in children after exposure to
organophosphates pesticides (OP) at low levels but possible mechanism is still unclear.

Objectives: We aimed at investigating the effects of prenatal exposure to OPs on inhibitory control of
10 to 12 year-old-children assessed by a motor inhibition task during functional magnetic resonance
imaging (fMRI).

Methods: Ninety-five children from the PELAGIE cohort (Brittany-France, from 2002) underwent an 31 32 fMRI examination during which inhibition was assessed by a Go/No-Go task. Task performance was 33 assessed by average response latency, commission rate and composite performance score (PS). 34 Whole brain activation was estimated by modeling the hemodynamic response related to inhibition 35 demand and successful inhibition. OP exposure was assessed by measuring six dialkylphosphate 36 (DAP) metabolites in the urine of women in early pregnancy (<19 WG). Concentrations were summed 37 to obtain overall levels of diethylphosphate (DE), dimethylphosphate (DM) and total non-specific 38 metabolites (DAP), standardized to homogenize sampling conditions and categorized into levels of 39 exposure: low (reference), moderate or high. Regression models were adjusted for potential 40 cofounders considered by restriction and statistical criteria.

41 Results: Moderate levels of DAP were associated with a decreased commission rate (β = -6.65%, p = 42 0.04), indicating improved performance. Increasing levels of DM and DE were associated with 43 decreased brain activity in the left inferior and bilateral superior frontal regions during successful 44 inhibition. We did not observe any differential activation related to inhibitory demands.

Discussion: These results suggest that prenatal OPs may be associated with altered pattern of brain
activity in regions related to inhibition among children and need to be confirmed by additional
studies.

48 Keywords :

49 prenatal exposure, organophosphates, neurodevelopment, motor inhibition, fMRI

ounderegion

50 Introduction

51 Organophosphate pesticides (OPs) have been used worldwide since the 1970's for agricultural and 52 pest control purposes due to their high acute neurotoxicity for insects and shorter environmental 53 persistence than that of the organochlorine insecticides they substituted. Their half-life is limited 54 from a few days to several months depending on the environmental conditions. However, they are 55 frequently found, as parent compounds or degradation products, in both outdoor and indoor French 56 environments (Coscollà et al. 2017; Raffy et al. 2017). A common exposure pathway of the general population is thought to be the ingestion of food or drinking water contaminated with 57 organophosphate residues. A study of the typical French diet among pregnant women in 2011 58 estimated the daily intake of OP residues (including chlorpyrifos) to be sufficient to present a 59 potential risk of neurochemical effects (de Gavelle et al. 2016). Other sources of exposure including 60 61 the use of household insecticides or living close to areas where OP insecticides are used for agriculture, are very likely in France (Raffy et al. 2017). Since the 2000's, various mixtures of OPs or 62 63 their metabolites have been frequently found in human biomonitoring studies, in particular in the urine or hair matrices of pregnant French women and children, suggesting widespread and repeated 64 OP exposure (Béranger et al. 2018; Cartier et al. 2016; Dereumeaux et al. 2016). 65

The primary neurotoxic property of organophosphate compounds for insect control is the inhibition of acetylcholinesterase (AChE) activity in synaptic junctions, resulting in hyper-excitability of the nervous system. Similar effects have also been described in humans after accidental acute exposure. Other neurotoxic mechanisms of OPs have then been described at low doses including noncholinergic effects, such as oxidative stress, neuro-inflammation, the alteration of axonal transport and mitochondria metabolism, the loss of neurons and glia, and persistent serotonergic effects (Chen et al. 2017; Slotkin and Seidler 2005; Terry 2012).

Low-dose exposure is a matter of concern, especially during prenatal and childhood periods. Biomonitoring studies have reported organophosphate metabolites to be found in the placenta, cord blood and meconium matrices, suggesting the ability of OPs to cross the placental barrier (Silver et al. 2017; Whyatt R M and Barr D B 2001). Given the vulnerability of the developing blood-brain-barrier especially during fetal life (Eskenazi et al. 1999; Gupta et al. 1999; Rice and Barone 2000), the epidemiological literature investigating potential adverse neurodevelopmental effects following prenatal exposure to OPs is extensive.

Although a variety of outcomes has been investigated, several studies have reported adverse effects on specific cognitive and behavioral functions that suggest possible alteration of executive functions (González-Alzaga et al. 2014; Koureas et al. 2012; Muñoz-Quezada et al. 2013). Most of published

83 studies used neuropsychological tests or standardized questionnaires. We assume that the 84 investigation of functional brain processes while children are engaged in inhibitory cognitive 85 processes, suspected to be affected by OPs, has the potential to provide mechanistic insights into 86 such processes. Task-related functional magnetic resonance imaging (fMRI) is one way to measure 87 cerebral activity during cognitive activity (Horton et al. 2014). A recently published pilot study used functional near-infrared spectroscopy (fNIRS) to investigate neural activity in Mexican-American 88 89 adolescents from the CHAMACOS cohort (Sagiv et al. 2019). FNIRS is based on the same technique as 90 fMRI, in which changes in neurovascular coupling during cerebral activity are detected. They reported associations between prenatal exposure to OP and altered activation pattern during tasks 91 92 involving executive functions and language comprehension. There was a bilaterally decreased activity 93 in the inferior frontal cortex during tasks of cognitive flexibility and working memory, which are two 94 executive functions.

95 Among executive functions, inhibitory control is critical for learning (Bari and Robbins 2013) and 96 difficulties in the ability to control and inhibit a prepotent response can lead to various 97 developmental challenges, such as ADHD (Willcutt et al. 2005) and poor academic performance 98 (Diamond et al. 2007). Our aim was to assess the association between prenatal exposure to OP 99 evaluated by maternal urinary biomarkers, and inhibitory control of 10- to 12- year-old children from 100 a population-based mother-child cohort using fMRI during a motor inhibition task.

101

102 Materials and Methods

103 Study Population

104 The PELAGIE (Perturbateurs endocriniens: Etude Longitudinale sur les Anomalies de la Grossesse, 105 l'Infertilité et l'Enfance) is a population-based mother-child cohort which included 3,421 women 106 before 19 weeks of gestation from three districts of Brittany (France) between 2002 et 2006 107 (Garlantézec et al. 2009). They were recruited by obstetricians, gynecologists, and ultra-108 sonographers at early visits for prenatal care. At inclusion, they completed a self-administered 109 questionnaire about family, social, and demographic characteristics, diet, and lifestyle and sent a 110 urine sample to the laboratory. At delivery, midwives and pediatricians at the maternity units 111 provided with information from medical records about the pregnancy, delivery, and neonatal health. Questionnaires given when their children were two and six years old allowed recording further 112 113 information about the lifestyle, development and health of the child.

A sub-cohort of 251 children between 10 and 12 years of age was randomly selected for the present
study to include approximatively 100 children. Children had to be born at term (delivery ≥ 35 weeks

116 of amenorrhea (WA)) and present no major condition at birth (neonatal hospitalization, 117 hypoglycemia, or five-minute Apgar score < 7), nor prenatal exposure to tobacco, alcohol, or medical 118 treatment during childhood which could affect neurodevelopment (methylphenidate, psychotropic, 119 or antiepileptic drugs, etc.) (Cartier et al. 2016). Among them, 124 (49.4%) families refused to participate or were lost to follow-up, 26 (10.3%) were excluded due to technical (braces) or medical 120 121 (meningitis, or head trauma requiring medical supervision) reasons. Thus, 101 (40.2%) children participated in the neuropsychological and fMRI examinations. Urine samples were not available for 122 123 six women because they were used for other urinary assays, resulting in 95 mother-child pairs for our 124 study population (inclusion flowchart is shown in Figure 1).

At the time of follow-up, parents completed a questionnaire about the environmental conditions and health of their child. All parent and child participants provided written informed consent and the appropriate ethics committees approved the study.

128 Organophosphate insecticide exposure assessment

129 At inclusion, women returned a urine sample (first morning void) in a 10-mL test tube (95 × 16-mm 130 polypropylene, with wing plug). Six nonspecific dialkylphosphate (DAP), metabolites of numerous 131 OPs analyzed: dimethylphosphate (DMP), dimethylthiophosphate (DMTP), were dimethyldithiophosphate (DMDTP), diethyl-phosphate (DEP), diethylthiophosphate (DETP), and 132 diethyldithiophosphate (DEDTP). Analyses were performed in 2007-2008 (n = 54), 2013 (n = 22), and 133 134 2017 (n = 19) by liquid chromatography-electrospray ionization tandem mass spectrometry (LC/ MS-135 MS) after solid-phase extraction (SPE) at the LABOCEA (Laboratoire public Conseil, Expertise et Analyse en Bretagne). Details of the chemical analysis procedures have been previously described 136 137 (Cartier et al. 2016).

The limits of quantification (LOQ) for the chemical analyses performed in 2007-2008 (hereafter 138 referred as series n°1) were 1.25, 1.7, 0.02, 0.2, 1, and 0.45 µg/L for DEP, DETP, DEDTP, DMP, DMTP, 139 140 and DMDTP, respectively. Values between the limit of detection (LOD) and the LOQ were available for series n°2 and n°3 analyzed in 2013 and 2017, respectively. The LODs were 0.2, 0.1, 0.005, 0.06, 141 0.32, and 0.13 µg/L for DEP, DETP, DEDTP, DMP, DMTP, and DMDTP, respectively. The mass 142 concentrations were converted to molar concentrations using the molecular masses (g/mol) from 143 144 Toxnet database: DEP (CAS number 598-02-7) 154.101, DETP (CAS number 2465-65-8) 170.168, DEDTP (CAS number 52857-42-8) 186.235, DMP (CAS number 813-78-5) 126.0473, DMTP (CAS 145 146 number 59401-04-6) 142.114, and DMDTP (CAS number 32534-66-0) 158.181. Concentrations were 147 summed to obtain overall concentrations of diethylphosphate metabolites (DE: DEP, DETP, DEDTP),

dimethylphosphate metabolites (DM: DMP, DMTP, DMDTP), and all nonspecific dialkylphosphatemetabolites (DAP).

150 Motor inhibition function

151 <u>The Go/No-Go paradigm</u>

The motor inhibition function of the children was evaluated using a 10-minute visual Go/No-Go task completed in the MR scanner. First, a short practice run in front of a computer and then in a mock MR scanner was performed to provide the instructions and accustom the child to the scanner environment.

156 The event-related task, adapted from Mostofsky et al. (2003), minimizes cognitive demands other 157 than motor execution and response inhibition and can be easily performed in children (Suskauer et 158 al. 2008). Green and red smileys were rapidly presented on a screen, every 1.5 s. Children were asked to press a button as quickly as possible when seeing a green smiley but had to refrain otherwise. The 159 160 ratio of No-go cues (red smileys) over Go cues (green smileys) was 1:4, to elicit a dominant motor response. The task was implemented using E-Prime v.2.0.8 Professional (Psychology Software Tools, 161 162 Pittsburgh, PA, USA) and presented using the Nordic Neurolab Solution (Nordic Neurolab, Bergen, 163 Norway).

The task was split into two runs of five minutes each, composed of 150 trials (123 Go and 27 No-Go cues) and four 10-second rest periods a run. A trial consisted of the consecutive presentation of a cue (duration of 200 ms) and a cross fixation point (duration of 1,300 ms). A response was allowed until the next trial began, giving a constant inter-stimulus interval of 1,500 ms. Smileys appeared in a pseudo-random order: cues at the beginning of runs or following resting periods had to be Go cues. Go cues had to occur at least three times in a row and No-Go cues had to occur at most two times in a row.

171 <u>Performance indicators at the Go/No-Go task</u>

Child performance was evaluated by response latency (average reaction time for the correct 172 173 answering of Go cues), commission rate (incorrect answers for the No-Go cues), and a composite 174 performance score. Responses occurring during cue presentation (before 200 ms) were considered 175 anticipatory and were excluded from the indicator calculation (on average, less than 1% of all cues). The accuracy of motor inhibition was assessed by the sensitivity index, d', which subtracts the 176 177 standardized (z score) commission rate from the standardized hit rate (correct answers for Go cues). 178 The response latencies and commission errors inversely correlated with each other (Spearman rho 179 coefficient =-0.31, p < 0.01). Thus, in order to combine both response speed and accuracy into a 180 single score, we built a composite performance score by subtracting the response latencies from the

accuracy of motor inhibition (d'), following standardization , based on the work of Collignon et al.2010.

Thus, children with the high scores were considered to efficiently perform the task (fast with a high hit rate, with few commissions), whereas children with the lowest scores were considered to be slow and found it difficult to inhibit their response to No-Go cues. Performance indicators were missing for two children (resulting in n = 93 included children), due to technical issues with the recording.

187 <u>Cerebral hemodynamic response related to the task</u>

Scanning was performed on a 3 T MR Scanner (Magnetom Verio, VB17, Siemens, Erlangen, Germany)
using a 32-channel receiver head coil.

High-resolution 3D anatomical images were obtained using T1-weighted MPRAGE at 1 mm³
resolution for anatomical referencing (repetition/inversion/echo times: 1,900/900/2.26 ms, 9° flip
angle).

Functional images were acquired using gradient echo-planar imaging (EPI) with repetition/echo times of 2,500/30 ms, a 90° flip angle, and 110 volumes per run. Each volume was composed of 34 axially oriented 4 mm-interleaved slices, covering the entire brain. Scans were based on a 110*110 acquisition matrix (220x220 mm² field of view), with a voxel size of 2*2*4 mm³. Operators assessed image quality during the acquisition and two children who moved repeated the sequence. Learning effects are not expected for this task and we did not observe any improvement in performance during the two sessions.

Image pre-processing is described in Supplementary Material. Head movements were evaluated by
six parameters calculated during the realignment step. Three children had head motion greater than
2 mm (in-plane size of an acquisition voxel) and they were excluded from further image analysis
(resulting in n = 92 included children).

204 The blood oxygen level dependent (BOLD) response was estimated from trials (Go cue and No-Go 205 cue) and motion regressors, with an implicit baseline corresponding to resting periods (enabling 206 recovery of the hemodynamic response) and the time between trials, using a canonical 207 hemodynamic response function (HRF) and its temporal derivative. This HRF is the subtraction of two 208 gamma functions, the first modelling the peak of intensity, with a latency of 6 s, and the second one 209 the undershoot during the recovery period, with a latency of 16 s. The use of a temporal derivative 210 allows for variations in peak latency, while providing comprehensive models for the response. 211 Scanner drift was modeled with a discrete cosine transform (DCT) set (128-second cut off) and 212 temporal autocorrelation was accounted for using an autoregressive AR(1) model over the entire

213 brain. First, we modeled Go cues and No-Go cues, independently of the subject's response. Secondly, 214 we modeled successful No-Go, Go, and failed No-Go conditions. Maps were extracted for each 215 condition and each subject by voxel-wise multiple linear regressions estimated by the restricted 216 maximum likelihood (ReML) method and then contrasted. We built two types of individual contrast images, as we were investigating the ability to stop a planned response when it is no longer 217 pertinent. First, we extracted "No-Go vs Go" activations, representing activation amplitudes that 218 were higher for inhibition than motor tasks, when children perceived the inhibition demand. 219 220 Secondly, we built contrast images for "Successful No-Go vs successful Go", when children were able to inhibit and stop their answers, expected to reflect successful inhibition. There were no differences 221 222 between the two runs for contrast estimates (at the uncorrected cluster level p < 0.001) or any performance indicators (at p < 0.05), allowing concatenation of the two runs. 223

224 Statistical analysis strategy

225 Standardization of sampling conditions and imputation

Sampling conditions of urinary samples of the present study are heterogeneous, in particular duration of storage due to three different chronological series of analyses (2007-2008, 2013 and 2017). Biomarker concentrations may vary according to sampling conditions. Thus, we aimed to standardize the urinary concentrations of DAP under various sampling conditions to limit their possible influence in the association study.

231 We adapted a standardization method based on regression residuals previously described by 232 Mortamais et al. (Mortamais et al. 2012). First, we estimated the influence of sampling conditions 233 (sampling day, duration of storage before and after freezing, and creatinine levels) using an adjusted 234 Tobit regression model to account for multiple censored data. We adjusted the Tobit models for the 235 year and season of sampling, as these parameters may be associated with the true level of exposure and reflect temporal variation of OP uses and exposure. Then, left-censored data (values below the 236 237 LOD) were imputed from a log-normal distribution (Jin et al. 2011). Last, imputed concentrations 238 were standardized using Tobit model estimates as if urines had been sampled under the same conditions (i.e., week-day collection, freezing before 24h after collection, duration of storage < 5 239 240 years and median creatinine concentration).

All OP metabolite concentrations used in further statistical analyses were imputed for values below the LOD or LOQ and standardized for sampling conditions (sampling day, duration of storage and creatinine levels) unless specified otherwise. DAP levels were categorized into three groups: for DAP and DM in terciles, and for DE (with > 33% of values < LOD), values < LOD, values > LOD, and < median, values > median. For clarity, we refer hereafter to low, moderate, or high levels of exposure.

246 <u>Statistical analyses</u>

247 Multivariable linear regression models were built on performance indicators (log-transformed-248 response latency, commission rate, and performance score) to analyze the association with each 249 categorical level of OP metabolite. We used restricted cubic splines with log-transformed metabolite 250 concentrations to assess a possible dose-response relation. When checking for the assumption of 251 linearity, we plotted urinary metabolite levels as continuous.

The BOLD response during the motor inhibition task was modeled by mixed effect generalized linear 252 regression models to investigate the effect-modification of the prenatal exposure to OP by the motor 253 254 inhibition task on cerebral activity. We did not have any a priori hypotheses on brain regions that 255 could be differentially activated during the inhibition task across exposure levels and thus performed 256 whole-brain analyses. We then concentrated our interpretation on the regions known to be part of 257 the motor inhibition network (frontal and anterior cingular cortices and supplementary motor area 258 (SMA)) regions (Mostofsky et al. 2003; Suskauer et al. 2008) and also consider other brain regions, such as those than can be involved in ancillary processes (e.g. attention, vision). Statistical 259 260 significance was assessed for cluster-wise significance, defined by random field theory (FWER 261 corrected p = 0.05) to account for 3D spatial autocorrelation, using an uncorrected one-sided cluster-262 defining threshold of p = 0.01 (Nichols and Hayasaka 2003).

Then, we applied our linear regression models to all the children stratified by sex to investigate whether the association between OP metabolite levels and motor inhibition differ according to the sex of the child.

266 <u>Potential cofounders</u>

267 Potential adjustment variables were selected from lifestyle and socioeconomic variables collected at 268 inclusion (i.e. maternal age, maternal educational level, pre-pregnancy BMI, pre-pregnancy tobacco consumption, fruit and vegetable consumption and fish consumption during pregnancy), from 269 270 medical records at delivery (i.e. pregnancy pathologies, child's sex, or during follow-up (i.e. duration 271 of non-exclusive breastfeeding, parity, child's educational level and child's lateralization at follow-272 up). They were included in all regression models if they were associated (at p < 0.10) with both OP 273 metabolite levels and motor inhibition performance (see Supplementary Material, Table S.3); only 274 breastfeeding (\leq vs > 3 months) met the criteria. Fruit and vegetable intake was also forced in all 275 regression models (< $vs \ge 3$ portions a day), as this has been described to be a possible source of DAP 276 exposure (Sudakin and Stone 2011).

277 <u>Sensitivity analyses</u>

278 Two sets of sensitivity analyses were performed. The first one excluded two children with extreme

- 279 performance scores (statistical outliers outside averaged performance score ± 3* standard errors
- 280 interval), and the second included OP metabolite levels imputed for values below the LOD but not

281 standardized for sampling conditions.

282 R software v.3.5.1 (https://www.R-project.org/) and SPM12 (http://www.fil.ion.ucl.ac.uk/spm/) were

used for data analysis.

284 <u>Results</u>

285 <u>Population characteristics</u>

286 All pregnant women included in the study were of European origin (Table 1). At inclusion, they had a 287 median age of 30.4 years, were highly educated (46.8% attended at least three years of post-288 secondary school) and declared to be employed (91.6%). Two thirds declared to be non smokers before inclusion. These women reported few medical problems during pregnancy (3.2% reported 289 290 high blood pressure, no diabetes). They gave birth mainly to girls (56.8%), with a median birth weight 291 of 3,430 g and 43.2% of the children were breastfed more than three months. Their children at the 292 follow-up time of the present study were mostly right-handed (87.4%), had a median age of 10.8 293 years, and most attended elementary school (72.6%). We did not observe any differences between 294 the study population and non-participants for these characteristics (see Supplementary material, 295 Table S.1).

296 Urinary OP metabolite levels

297 OP metabolites were detected in the urine of 57% of the pregnant women for DE, 86% for DM and 298 95% for all DAP compounds with respective median values of 0.8, 31.8, and 44.4 μ M (see 299 Supplementary material, Table S.2).

300 Increased creatinine concentrations were associated with increased levels of DE metabolites (1.10⁻⁴ μ M for an increase of one μ M of creatinine, p = 0.02). A moderate duration of storage before 301 302 freezing (24-48 h) versus a short (< 24h) duration was associated with increased DM metabolite levels (p < 0.01) and a moderate duration of storage after freezing (5-10 years) versus a short (< 5 303 304 years) duration was associated with decreased DM metabolite concentrations (p < 0.01). We found 305 no statistically significant association between sampling conditions and DAP metabolite levels nor 306 difference with sampling day (weekday vs weekend) for any OP metabolite (see Supplementary 307 Material, Table S.2). There was an association between analyses series and DAP metabolites ($\beta = -1.0$ 308 for series 2 vs series 1 and β = 0.3 for series 3 vs series 1, global p = 0.03), for DM metabolites (β = -2.6 for series 2 vs series 1 and β = 0.3 for series 3 vs series 1, global p < 0.001) and not for DE 309

- 310 metabolites (β = 0.06 for series 2 vs series 1 and β = -0.34 for series 3 vs series 1, global p = 0.91) (see
- 311 Supplementary Material Figure S.I and Table S.2).
- 312 As expected, differences did not remain statistically significant after standardization of sampling

313 conditions (for DAP, p = 0.82; for DM, p = 0.23) (see Supplementary Material Figure S.I). The median

- standardized OP metabolite levels presented were 0.8 μM, 17μM, and 25 μM for DE, DM, and DAP,
- 315 respectively (see Supplementary Material Table S.3).
- 316 Performance indicators and brain activation during the task

Children performed the Go/No-Go task with a median commission rate of 25.9%, and reaction time of 399.6 ms. The performance score ranged from -9.8 to 2.1, with a median of 0.3 (see Supplementary Material Table S.3).

For "No-go vs Go" and "Successful No-Go vs successful Go" contrasts in the brain, children showed activation in regions known to be involved in the motor inhibition network: the anterior cingulate/SMA and inferior and middle frontal regions of both hemispheres. These contrasts were also associated with activated clusters in the right middle temporal region and brain structures of both hemispheres (occipital and parietal region, posterior lobe of the cerebellum, caudate nucleus, and putamen) (see Supplementary Material, Table S.4 and Table S.5).

326 Prenatal urinary organophosphate metabolite levels and motor inhibition

327 There was an association between moderate levels of DAP and decreased commission rate (β = -328 6.65%, 95% confidence interval (CI): -12.83; -0.48), without a statistically significantly higher performance score (β = 0.33, CI: -0.49; 1.15). There was a trend towards a lower performance score 329 $(\beta = -0.76, CI: -1.56; 0.03)$ and longer response latency $(\beta = (e^{0.08}-1)*100 = 8\%, CI: 0; 16)$ associated 330 with moderate DE levels. Children whose mothers had moderate urinary levels of DM metabolites 331 332 had a lower commission rate (β = -5.59), but the association was not statistically significant (CI: -333 11.71; 0.53) (Figure 2). Our results did not show a linear dose-response relation between any child's inhibition performance indicators and maternal urinary OP metabolite levels when OP metabolite 334 concentrations were used as continuous variables (Figure 2 and Supplementary material Figure S.II). 335 The associations were similar for children of either sex, with no difference in the estimates, except 336 337 for commission rate and moderate DE levels (β = 8.23, CI: -0.17; 16.63 for girls (n = 53) and β = -7.33, CI: -15.98; 1.33 for boys (n = 40)). 338

The first set of sensitivity analyses, which excluded statistical outliers, confirmed the previous statistically significant associations observed between commission rate and moderate DAP levels. Most of the associations were similar in the second set of sensitivity analyses when the OP metabolite levels were not standardized for sampling condition, except that response latency was

significantly higher for children whose mothers showed high DE urinary levels (β = 8%, CI: 0 ; 16) (see Supplementary material Tables S.4 and S.5).

We did not observe any differential activation related to inhibition demand perception. We did not observe any difference in the inhibition network, stratified by sex, except for girls, with decreased brain activity in the medial superior frontal region in association with moderate levels of DM metabolites see Supplementary Material, Table S.6).

When children succeeded in inhibiting their response, there was decreased activation in the left inferior frontal regions associated with high urinary levels of DM metabolites. Children whose mothers had moderate or high levels of DE metabolites exhibited a lower BOLD response in the left and right frontal regions (Figure 3 and Table 2). We did not observe any differential brain activation in girls when stratifying the analyses by sex. However, we observed decreased brain activity in the frontal regions of boys whose mothers had high levels of DAP or DM metabolites or moderate urinary levels of DE metabolites.

356 Discussion

The present study investigated the potential neurodevelopmental toxicity of prenatal exposure to OPs on motor inhibition function in a Brittany population-based cohort of children. We did not observe any major alterations of the children's performance on the Go/No-Go task. A decreased commission rate associated with moderate levels of DAP suggests better performance. There was no differential brain activation related to inhibitory demands, but we observed decreased cerebral activity related to successful inhibition in regions involved in the motor inhibition network.

363 Previous epidemiological studies investigating the neurodevelopmental effects of prenatal exposure to OP have suggested possible alterations in children's inhibitory control. In a recent pilot study, 364 Sagiv et al. reported an association between prenatal exposure to OPs and altered patterns of brain 365 366 activity during tasks of executive function and language comprehension using the fNIRS technique on 367 95 adolescents from a population of agricultural communities of the Salinas Valley in California (CHAMACOS study). They found null associations during the Go/No-Go task but reported decreased 368 369 bilateral brain activation in the prefrontal cortex during a cognitive flexibility task (Sagiv et al. 2019). In the CHAMACOS study, prenatal exposure to OP was also assessed by two maternal urinary 370 samples collected during the first and second halves of pregnancy. The mothers had higher levels of 371 DAP (109 nM DAP, 17.7 nM DE and 76.8 nM DM) than the present study. The authors reported 372 373 statistically greater attention problems, ADHD, and K-CPT ADHD Confidence Index scores for five-374 year-old children (n = 323) and a similar but non-significant association at 3.5 years of age (Marks et 375 al. 2010). Furlong et al. investigated the association between prenatal exposure to OP and

376 neurodevelopmental phenotypes in six- to nine-year-old children (n = 141) at the Mount Sinai 377 Children's Environmental Health Center. Exposure was assessed by analyzing maternal urine 378 collected between 25 and 40 weeks of gestation. Urinary levels were higher than in the PELAGIE 379 cohort (median of 37.1 nM DM and 16.6 nM DE). They reported positive associations between DM metabolite levels and executive functioning factor scores. The impulsivity and externalizing factor 380 score was not associated with maternal urinary levels of DE or DM (Furlong et al. 2017). A cross-381 382 sectional study investigated the effect of prenatal and postnatal exposure to OP insecticides on the 383 neurodevelopment of six- to eight-year-old Ecuadorian primary school children (n = 84) and reported a borderline significant association between reaction time and attention test scores and the 384 385 children's current exposure (Harari Raul et al. 2010). A cross-sectional study evaluated the effects of 386 methyl-parathion (MP) exposure during childhood on six-year-old-children living in Mississippi and 387 Ohio. Exposed children (n=132) were more prone to attention, and behavioral skill problems than 388 controls (n = 147). However, the results were not consistent between the two sites of the study 389 (Ruckart et al. 2004). Although they were not specifically investigating inhibition function, the results 390 of Rauh et al. showing that seven-year-old children with high prenatal exposure to chlorpyrifos had frontal cortical thinning, also suggest brain alterations in the frontal cortex, involved in inhibition 391 392 (Rauh et al. 2012). Findings of brain alterations after prenatal exposure to OPs are supported by one 393 animal study that exposed gestational dams to 5mg/kg/day chlorpyrifos during gestational days 13 to 394 17. At postnatal day 60, the authors reported reduced neuron and glia count in anterior cingulate, 395 prelimbic, and infralimbic areas of medial prefrontal cortex (mPFC), a region of the inhibition 396 network, at levels showing no effect on body weight, organ coefficient, or pathological morphological 397 changes (such as cytoplasm swelling, enlarged intercellular spaces, or cell apoptosis) in the mPFC 398 (Chen et al. 2017).

399 Two major types of inhibitory control can be distinguished, cognitive and behavioral inhibition (Bari 400 and Robbins 2013). The task we used involves behavioral response inhibition with a motor 401 component. Response inhibition appears in early childhood and matures until adolescence (Luna and Sweeney 2006). Performance indicators in our population of 10- to 12-year-old children were similar 402 403 to those reported during a similar Go/No-Go task with 8- to 13-year-old children (median response 404 latency of 400 ms vs 407 ms and 25.9% vs 22% for commission errors) (Suskauer et al. 2008). We did 405 not observe any differential performance score for the Go/No-Go task but observed decreased brain 406 activity in frontal regions, known to be involved in inhibitory control. Children who are the most 407 highly exposed to OP during pregnancy may present compensatory brain functioning to prevent problems of inhibition. However, we cannot eliminate the possibility that our neuropsychological 408 409 indicators did not have sufficient sensitivity to discriminate the inhibitory capacity between the

children. In the same way, we cannot rule out that our study (n=95) lacks of statistical power to point
out any association, in particular with the sex-stratified analyses (n=41 boys/n=54 girls).

412 The motor inhibition network concerns three main regions: the anterior cingulate, SMA, and inferior 413 frontal gyri of both hemispheres (Aron 2007; Mostofsky et al. 2003). As expected, we observed 414 cerebral activation in these regions in the contrasts that we explored. We also observed additional 415 activation in other brain structures of both hemispheres (occipital and parietal region, posterior lobe 416 of the cerebellum, caudate nucleus and putamen). These regions are not part of the motor inhibition 417 network but have been described to be involved in certain motor inhibition tasks, in particular in 8-418 to 13-year-old children (Suskauer et al. 2008). Activation observed in these brain areas may reflect 419 ancillary processes necessary for the Go/No-Go task, such as vision processing, attention and 420 decision making (Behrmann et al. 2004; Grahn et al. 2008; Rosen et al. 2018; Stoodley and 421 Schmahmann 2009). Nevertheless, BOLD imaging measures the ratio of deoxyhemoglobin to 422 oxyhemoglobin to indirectly evaluate neural activity. This indicator is commonly used, but may vary 423 depending on individual characteristics (blood flow, iron deficiency, etc.) and the performance of 424 inhibition is not determined solely by the intensity of the hemodynamic response but could also be 425 explained by brain structure, connectivity, etc. (Ogawa et al. 1990). Reduced activation may indicate 426 that exposure altered the recruitment of neural resources, such as a lack of neural fibers and/or 427 reduced connectivity of the inhibitory network. The associations observed between prenatal levels of DM and DE metabolites and decreased activity from different regions of the brain may suggest an 428 429 alteration in the pattern of brain activity.

430 We observed no sex-specific effects on performance indicators but observed decreased brain activity 431 related to successful inhibition only in boys. This finding is consistent with the literature, which 432 suggests that boys may be more sensitive to OP exposure (Comfort and Re 2017; Marks et al. 2010). 433 In their review, Comfort and Re reported that sex-specific effects of OP exposure have been 434 commonly found in rodent studies. Whereas, the exact mechanisms of sex-dependent OP effects are not fully established, authors assumed it could be the combination of altered morphology, 435 436 neuropeptide and neurotransmitter signaling, and neuroinflammation. They highlighted the 437 necessity to investigate this effect in epidemiological studies. Indeed, Marks et al. suggested an increased risk of ADHD at five years of age in the CHAMACOS cohort in association with prenatal 438 439 exposure to OP in boys only.

Our study was based on a mother-child cohort (PELAGIE study) in which the longitudinal design
provides the opportunity to measure OP insecticide exposure during early pregnancy. We were also
able to consider lifestyle and socioeconomic factors by restriction or adjustment. In particular,

443 children whose mothers reported alcohol or tobacco consumption during pregnancy were not included in the study to eliminate two established risk factors of poorer neurodevelopment or 444 445 behavioral troubles in children (Yolton et al. 2014). We assessed prenatal exposure to OPs by 446 maternal urinary biomarkers, which gives us the ability to capture all sources of exposure. We used 447 fMRI to investigate brain functioning during an inhibition task in association with prenatal exposure to OP in the general population. This technique provides excellent spatial resolution, making it 448 449 possible to investigate the potential effects of exposure in brain regions of only a few cubic 450 millimeters.

451 Although we did not find any differences in the characteristics of our study population and non-452 participants, limiting the risk that a confounder may predict participation, the design of the PELAGIE 453 cohort does not prevent a possible selection bias. Unlike the previously described epidemiological 454 studies, the PELAGIE population does not include an ethnic minority and consists of participants with 455 a high socioeconomic status. We cannot rule out the possibility that the stimulating environments 456 and possible healthier lifestyles in our study population attenuated the potential neurotoxicity of OP 457 exposure. The main limitation of our study was the exposure assessment, based on only one (first) 458 morning void urinary sample collected during the first trimester of pregnancy, whereas OP 459 insecticides have a short half-life, from several hours to several days (Costa 2018). It is likely that our 460 measurement did not capture intra-individual variability, over time, due to the rapid metabolism and excretion of OP insecticides. This produces non-differential measurement errors in the exposure 461 462 estimation. We could not explore the effects of OP exposure during late pregnancy or childhood, 463 although human neurodevelopment is known to continue during childhood, until adolescence for 464 maturation of the prefrontal cortex (Rice and Barone 2000). A recent study in the Netherlands, 465 showing few effects of prenatal exposure to OP on nonverbal IQ in six-year-old children, suggests 466 that late pregnancy may be a susceptible window for adverse effects on cognition (Jusko et al. 2019). 467 DAP metabolites are found in food and dust. Thus, measurements of urinary levels of DAP may 468 reflect intake of both the toxic parent OP compound and the metabolite itself, without possible distinction and biased exposure classification (Sudakin and Stone 2011). 469

470 Conclusion

We found no major alteration of neuropsychological scores in a French population-based cohort of 10 to 12 year-old-children with a high socioeconomic status, but found decreased brain activity in areas related to inhibition in association with prenatal exposure to OPs. These findings need to be confirmed by other studies and more deeply investigated by further neuroimaging techniques of morphometry, and functional or structural connectivity.

476 Acknowledgements

We are grateful to the gynecologists, obstetricians, ultra-sonographers, midwives, pediatricians, and families who participated in the study. We particularly thank Catherine Ducassoux, the MR technicians, and the clinical investigation unit of Rennes hospital for neuropsychological and MRI testing.

This study is supported by the Fondation de France (FdF) and Rennes University Hospital. The PELAGIE cohort is funded by Inserm, the French Ministries of Health, Labor, and Ecology, InVS, ANR, and ANSES. MRI data acquisition was performed at the Neurinfo MRI research facility of the University of Rennes I. Neurinfo is funded by the European Union (FEDER), the French State, the Brittany Council, Rennes Metropole, Inria, Inserm, and the University Hospital of Rennes.

486 <u>References</u>

- 487 Aron AR. 2007. The Neural Basis of Inhibition in Cognitive Control. The Neuroscientist 13:214–228;
 488 doi:10.1177/1073858407299288.
- Bari A, Robbins TW. 2013. Inhibition and impulsivity: behavioral and neural basis of response control.
 Progress in Neurobiology 108:44–79; doi:10.1016/j.pneurobio.2013.06.005.
- 491Behrmann M, Geng JJ, Shomstein S. 2004. Parietal cortex and attention. Current Opinion in492Neurobiology 14:212–217; doi:10.1016/j.conb.2004.03.012.
- 493Béranger R, Hardy EM, Dexet C, Guldner L, Zaros C, Nougadère A, et al. 2018. Multiple pesticide494analysis in hair samples of pregnant French women: Results from the ELFE national birth495cohort. Environment International 120:43–53; doi:10.1016/j.envint.2018.07.023.
- 496 Cartier C, Warembourg C, Le Maner-Idrissi G, Lacroix A, Rouget F, Monfort C, et al. 2016.
 497 Organophosphate Insecticide Metabolites in Prenatal and Childhood Urine Samples and
 498 Intelligence Scores at 6 Years of Age: Results from the Mother–Child PELAGIE Cohort
 499 (France). Environ Health Perspect 124:674–680; doi:10.1289/ehp.1409472.
- Chen XP, Chao YS, Chen WZ, Dong JY. 2017. Mother gestational exposure to organophosphorus
 pesticide induces neuron and glia loss in daughter adult brain. Journal of Environmental
 Science and Health, Part B 52:77–83; doi:10.1080/03601234.2016.1239973.
- Collignon O, Girard S, Gosselin F, Saint-Amour D, Lepore F, Lassonde M. 2010. Women process
 multisensory emotion expressions more efficiently than men. Neuropsychologia 48:220–225;
 doi:10.1016/j.neuropsychologia.2009.09.007.
- 506Comfort N, Re DB. 2017. Sex-specific Neurotoxic Effects of Organophosphate Pesticides Across the507Life Course. Curr Environ Health Rep 4:392–404; doi:10.1007/s40572-017-0171-y.
- Coscollà C, López A, Yahyaoui A, Colin P, Robin C, Poinsignon Q, et al. 2017. Human exposure and risk
 assessment to airborne pesticides in a rural French community. Science of The Total
 Environment 584–585:856–868; doi:10.1016/j.scitotenv.2017.01.132.
- 511 Costa LG. 2018. Organophosphorus Compounds at 80: Some Old and New Issues. Toxicol Sci 162:24– 512 35; doi:10.1093/toxsci/kfx266.
- 513de Gavelle E, de Lauzon-Guillain B, Charles M-A, Chevrier C, Hulin M, Sirot V, et al. 2016. Chronic514dietary exposure to pesticide residues and associated risk in the French ELFE cohort of515pregnant women. Environment International 92–93:533–542;516doi:10.1016/j.envint.2016.04.007.
- 517 Dereumeaux C, Saoudi A, Pecheux M, Berat B, de Crouy-Chanel P, Zaros C, et al. 2016. Biomarkers of
 518 exposure to environmental contaminants in French pregnant women from the Elfe cohort in
 519 2011. Environment International 97:56–67; doi:10.1016/j.envint.2016.10.013.
- Diamond A, Barnett WS, Thomas J, Munro S. 2007. Preschool Program Improves Cognitive Control.
 Science 318:1387–1388; doi:10.1126/science.1151148.
- Eskenazi B, Bradman A, Castorina R. 1999. Exposures of children to organophosphate pesticides and
 their potential adverse health effects. Environmental Health Perspectives 107:409–419;
 doi:10.1289/ehp.99107s3409.

- Furlong MA, Herring A, Buckley JP, Goldman BD, Daniels JL, Engel LS, et al. 2017. Prenatal exposure to
 organophosphorus pesticides and childhood neurodevelopmental phenotypes.
 Environmental Research 158:737–747; doi:10.1016/j.envres.2017.07.023.
- Garlantézec R, Monfort C, Rouget F, Cordier S. 2009. Maternal occupational exposure to solvents and
 congenital malformations: a prospective study in the general population. Occupational and
 Environmental Medicine 66:456–463; doi:10.1136/oem.2008.041772.
- 531González-Alzaga B, Lacasaña M, Aguilar-Garduño C, Rodríguez-Barranco M, Ballester F, Rebagliato M,532et al. 2014. A systematic review of neurodevelopmental effects of prenatal and postnatal533organophosphatepesticideexposure.ToxicologyLetters230:104–121;534doi:10.1016/j.toxlet.2013.11.019.
- Grahn JA, Parkinson JA, Owen AM. 2008. The cognitive functions of the caudate nucleus. Progress in
 Neurobiology 86:141–155; doi:10.1016/j.pneurobio.2008.09.004.
- 537Gupta A, Agarwal R, Shukla GS. 1999. Functional impairment of blood-brain barrier following538pesticide exposureduring early development in rats. Hum Exp Toxicol 18:174–179;539doi:10.1177/096032719901800307.
- Harari Raul, Julvez Jordi, Murata Katsuyuki, Barr Dana, Bellinger David C., Debes Frodi, et al. 2010.
 Neurobehavioral Deficits and Increased Blood Pressure in School-Age Children Prenatally
 Exposed to Pesticides. Environmental Health Perspectives 118:890–896;
 doi:10.1289/ehp.0901582.
- Horton MK, Margolis AE, Tang C, Wright R. 2014. Neuroimaging is a novel tool to understand the
 impact of environmental chemicals on neurodevelopment. Curr Opin Pediatr 26:230–236;
 doi:10.1097/MOP.0000000000074.
- Jin Y, Hein MJ, Deddens JA, Hines CJ. 2011. Analysis of Lognormally Distributed Exposure Data with
 Repeated Measures and Values below the Limit of Detection Using SAS. Ann Occup Hyg
 55:97–112; doi:10.1093/annhyg/meq061.
- Jusko TA, van den Dries MA, Pronk A, Shaw Pamela A., Guxens Mònica, Spaan Suzanne, et al. 2019.
 Organophosphate Pesticide Metabolite Concentrations in Urine during Pregnancy and
 Offspring Nonverbal IQ at Age 6 Years. Environmental Health Perspectives 127:017007;
 doi:10.1289/EHP3024.
- Koureas M, Tsakalof A, Tsatsakis A, Hadjichristodoulou C. 2012. Systematic review of biomonitoring
 studies to determine the association between exposure to organophosphorus and pyrethroid
 insecticides and human health outcomes. Toxicology Letters 210:155–168;
 doi:10.1016/j.toxlet.2011.10.007.
- Luna B, Sweeney JA. 2006. The Emergence of Collaborative Brain Function: fMRI Studies of the
 Development of Response Inhibition. Annals of the New York Academy of Sciences
 1021:296–309; doi:10.1196/annals.1308.035.
- Marks AR, Harley Kim, Bradman Asa, Kogut Katherine, Barr Dana Boyd, Johnson Caroline, et al. 2010.
 Organophosphate Pesticide Exposure and Attention in Young Mexican-American Children:
 The CHAMACOS Study. Environmental Health Perspectives 118:1768–1774;
 doi:10.1289/ehp.1002056.

- Mortamais M, Chevrier C, Philippat C, Petit C, Calafat AM, Ye X, et al. 2012. Correcting for the
 influence of sampling conditions on biomarkers of exposure to phenols and phthalates: a 2step standardization method based on regression residuals. Environ Health 11:29;
 doi:10.1186/1476-069X-11-29.
- Mostofsky SH, Schafer JG, Abrams MT, Goldberg MC, Flower AA, Boyce A, et al. 2003. fMRI evidence
 that the neural basis of response inhibition is task-dependent. Cognitive brain research 17:
 419–430.
- 572 Muñoz-Quezada MT, Lucero BA, Barr DB, Steenland K, Levy K, Ryan PB, et al. 2013.
 573 Neurodevelopmental effects in children associated with exposure to organophosphate
 574 pesticides: A systematic review. Neurotoxicology 39:158–168;
 575 doi:10.1016/j.neuro.2013.09.003.
- 576Nichols T, Hayasaka S. 2003. Controlling the familywise error rate in functional neuroimaging: a577comparative review. Statistical Methods in Medical Research 12:419–446;578doi:10.1191/0962280203sm341ra.
- 579 Ogawa S, Lee TM, Kay AR, Tank DW. 1990. Brain magnetic resonance imaging with contrast 580 dependent on blood oxygenation. Proc Natl Acad Sci U S A 87: 9868–9872.
- Raffy G, Mercier F, Blanchard O, Derbez M, Dassonville C, Bonvallot N, et al. 2017. Semi-volatile
 organic compounds in the air and dust of 30 French schools: a pilot study. Indoor Air 27:114–
 127; doi:10.1111/ina.12288.
- Rauh VA, Perera FP, Horton MK, Whyatt RM, Bansal R, Hao X, et al. 2012. Brain anomalies in children
 exposed prenatally to a common organophosphate pesticide. Proceedings of the National
 Academy of Sciences 109:7871–7876; doi:10.1073/pnas.1203396109.
- 587Rice D, Barone S. 2000. Critical periods of vulnerability for the developing nervous system: evidence588from humans and animal models. Environ Health Perspect 108:511–533;589doi:10.1289/ehp.00108s3511.
- Rosen ML, Sheridan MA, Sambrook KA, Peverill MR, Meltzoff AN, McLaughlin KA. 2018. The Role of
 Visual Association Cortex in Associative Memory Formation across Development. J Cogn
 Neurosci 30:365–380; doi:10.1162/jocn_a_01202.
- Ruckart P, Kakolewski Kirsten, Bove Frank J, Kaye Wendy E. 2004. Long-term neurobehavioral health
 effects of methyl parathion exposure in children in Mississippi and Ohio. Environmental
 Health Perspectives 112:46–51; doi:10.1289/ehp.6430.
- Sagiv SK, Bruno JL, Baker JM, Palzes V, Kogut K, Rauch S, et al. 2019. Prenatal exposure to organophosphate pesticides and functional neuroimaging in adolescents living in proximity to pesticide application. PNAS 201903940; doi:10.1073/pnas.1903940116.
- Silver MK, Shao J, Zhu B, Chen M, Xia Y, Kaciroti N, et al. 2017. Prenatal naled and chlorpyrifos
 exposure is associated with deficits in infant motor function in a cohort of Chinese infants.
 Environment International 106:248–256; doi:10.1016/j.envint.2017.05.015.
- Slotkin TA, Seidler FJ. 2005. The alterations in CNS serotonergic mechanisms caused by neonatal
 chlorpyrifos exposure are permanent. Developmental Brain Research 158:115–119;
 doi:10.1016/j.devbrainres.2005.06.008.

605Stoodley CJ, Schmahmann JD. 2009. Functional topography in the human cerebellum: A meta-606analysisofneuroimagingstudies.NeuroImage44:489–501;607doi:10.1016/j.neuroimage.2008.08.039.

Sudakin DL, Stone DL. 2011. Dialkyl phosphates as biomarkers of organophosphates: The current
 divide between epidemiology and clinical toxicology. Clinical Toxicology 49:771–781;
 doi:10.3109/15563650.2011.624101.

Suskauer SJ, Simmonds DJ, Fotedar S, Blankner JG, Pekar JJ, Denckla MB, et al. 2008. Functional
 magnetic resonance imaging evidence for abnormalities in response selection in attention
 deficit hyperactivity disorder: differences in activation associated with response inhibition
 but not habitual motor response. J Cogn Neurosci 20:478–493; doi:10.1162/jocn.2008.20032.

615Terry AV. 2012. Functional consequences of repeated organophosphate exposure: Potential non-616cholinergic mechanisms. Pharmacology & Therapeutics 134:355–365;617doi:10.1016/j.pharmthera.2012.03.001.

618 Whyatt R M, Barr D B. 2001. Measurement of organophosphate metabolites in postpartum
619 meconium as a potential biomarker of prenatal exposure: a validation study. Environmental
620 Health Perspectives 109:417–420; doi:10.1289/ehp.01109417.

Willcutt EG, Doyle AE, Nigg JT, Faraone SV, Pennington BF. 2005. Validity of the Executive Function
 Theory of Attention-Deficit/Hyperactivity Disorder: A Meta-Analytic Review. Biological
 Psychiatry 57:1336–1346; doi:10.1016/j.biopsych.2005.02.006.

Yolton K, Cornelius M, Ornoy A, McGough J, Makris S, Schantz S. 2014. Exposure to neurotoxicants
and the development of attention deficit hyperactivity disorder and its related behaviors in
childhood. Neurotoxicology and Teratology 44:30–45; doi:10.1016/j.ntt.2014.05.003.

627

- Bari A, Robbins TW. 2013. Inhibition and impulsivity: behavioral and neural basis of response control.
 Progress in Neurobiology 108:44–79; doi:10.1016/j.pneurobio.2013.06.005.
- 630 Behrmann M, Geng JJ, Shomstein S. 2004. Parietal cortex and attention. Current Opinion in 631 Neurobiology 14:212–217; doi:10.1016/j.conb.2004.03.012.

Béranger R, Hardy EM, Dexet C, Guldner L, Zaros C, Nougadère A, et al. 2018. Multiple pesticide
analysis in hair samples of pregnant French women: Results from the ELFE national birth
cohort. Environment International 120:43–53; doi:10.1016/j.envint.2018.07.023.

635 Cartier C, Warembourg C, Le Maner-Idrissi G, Lacroix A, Rouget F, Monfort C, et al. 2016.
636 Organophosphate Insecticide Metabolites in Prenatal and Childhood Urine Samples and
637 Intelligence Scores at 6 Years of Age: Results from the Mother–Child PELAGIE Cohort
638 (France). Environ Health Perspect 124:674–680; doi:10.1289/ehp.1409472.

639 Chen XP, Chao YS, Chen WZ, Dong JY. 2017. Mother gestational exposure to organophosphorus
640 pesticide induces neuron and glia loss in daughter adult brain. Journal of Environmental
641 Science and Health, Part B 52:77–83; doi:10.1080/03601234.2016.1239973.

Collignon O, Girard S, Gosselin F, Saint-Amour D, Lepore F, Lassonde M. 2010. Women process
 multisensory emotion expressions more efficiently than men. Neuropsychologia 48:220–225;
 doi:10.1016/j.neuropsychologia.2009.09.007.

- Comfort N, Re DB. 2017. Sex-specific Neurotoxic Effects of Organophosphate Pesticides Across the
 Life Course. Curr Environ Health Rep 4:392–404; doi:10.1007/s40572-017-0171-y.
- Coscollà C, López A, Yahyaoui A, Colin P, Robin C, Poinsignon Q, et al. 2017. Human exposure and risk
 assessment to airborne pesticides in a rural French community. Science of The Total
 Environment 584–585:856–868; doi:10.1016/j.scitotenv.2017.01.132.
- Costa LG. 2018. Organophosphorus Compounds at 80: Some Old and New Issues. Toxicol Sci 162:24–
 35; doi:10.1093/toxsci/kfx266.
- 652de Gavelle E, de Lauzon-Guillain B, Charles M-A, Chevrier C, Hulin M, Sirot V, et al. 2016. Chronic653dietary exposure to pesticide residues and associated risk in the French ELFE cohort of654pregnantwomen.655doi:10.1016/j.envint.2016.04.007.
- Dereumeaux C, Saoudi A, Pecheux M, Berat B, de Crouy-Chanel P, Zaros C, et al. 2016. Biomarkers of
 exposure to environmental contaminants in French pregnant women from the Elfe cohort in
 2011. Environment International 97:56–67; doi:10.1016/j.envint.2016.10.013.
- Diamond A, Barnett WS, Thomas J, Munro S. 2007. Preschool Program Improves Cognitive Control.
 Science 318:1387–1388; doi:10.1126/science.1151148.
- Eskenazi B, Bradman A, Castorina R. 1999. Exposures of children to organophosphate pesticides and
 their potential adverse health effects. Environmental Health Perspectives 107:409–419;
 doi:10.1289/ehp.99107s3409.
- Furlong MA, Herring A, Buckley JP, Goldman BD, Daniels JL, Engel LS, et al. 2017. Prenatal exposure to
 organophosphorus pesticides and childhood neurodevelopmental phenotypes.
 Environmental Research 158:737–747; doi:10.1016/j.envres.2017.07.023.
- Garlantézec R, Monfort C, Rouget F, Cordier S. 2009. Maternal occupational exposure to solvents and
 congenital malformations: a prospective study in the general population. Occupational and
 Environmental Medicine 66:456–463; doi:10.1136/oem.2008.041772.
- Grahn JA, Parkinson JA, Owen AM. 2008. The cognitive functions of the caudate nucleus. Progress in
 Neurobiology 86:141–155; doi:10.1016/j.pneurobio.2008.09.004.
- Gupta A, Agarwal R, Shukla GS. 1999. Functional impairment of blood-brain barrier following
 pesticide exposure during early development in rats. Hum Exp Toxicol 18:174–179;
 doi:10.1177/096032719901800307.
- Harari Raul, Julvez Jordi, Murata Katsuyuki, Barr Dana, Bellinger David C., Debes Frodi, et al. 2010.
 Neurobehavioral Deficits and Increased Blood Pressure in School-Age Children Prenatally
 Exposed to Pesticides. Environmental Health Perspectives 118:890–896;
 doi:10.1289/ehp.0901582.
- Horton MK, Margolis AE, Tang C, Wright R. 2014. Neuroimaging is a novel tool to understand the
 impact of environmental chemicals on neurodevelopment. Curr Opin Pediatr 26:230–236;
 doi:10.1097/MOP.0000000000074.
- Jin Y, Hein MJ, Deddens JA, Hines CJ. 2011. Analysis of Lognormally Distributed Exposure Data with
 Repeated Measures and Values below the Limit of Detection Using SAS. Ann Occup Hyg
 55:97–112; doi:10.1093/annhyg/meq061.

- Jusko TA, van den Dries MA, Pronk A, Shaw Pamela A., Guxens Mònica, Spaan Suzanne, et al. 2019.
 Organophosphate Pesticide Metabolite Concentrations in Urine during Pregnancy and
 Offspring Nonverbal IQ at Age 6 Years. Environmental Health Perspectives 127:017007;
 doi:10.1289/EHP3024.
- Luna B, Sweeney JA. 2006. The Emergence of Collaborative Brain Function: fMRI Studies of the
 Development of Response Inhibition. Annals of the New York Academy of Sciences
 1021:296–309; doi:10.1196/annals.1308.035.
- Marks AR, Harley Kim, Bradman Asa, Kogut Katherine, Barr Dana Boyd, Johnson Caroline, et al. 2010.
 Organophosphate Pesticide Exposure and Attention in Young Mexican-American Children:
 The CHAMACOS Study. Environmental Health Perspectives 118:1768–1774;
 doi:10.1289/ehp.1002056.
- Mortamais M, Chevrier C, Philippat C, Petit C, Calafat AM, Ye X, et al. 2012. Correcting for the
 influence of sampling conditions on biomarkers of exposure to phenols and phthalates: a 2step standardization method based on regression residuals. Environ Health 11:29;
 doi:10.1186/1476-069X-11-29.
- Mostofsky SH, Schafer JG, Abrams MT, Goldberg MC, Flower AA, Boyce A, et al. 2003. fMRI evidence
 that the neural basis of response inhibition is task-dependent. Cognitive brain research 17:
 419–430.
- Nichols T, Hayasaka S. 2003. Controlling the familywise error rate in functional neuroimaging: a
 comparative review. Statistical Methods in Medical Research 12:419–446;
 doi:10.1191/0962280203sm341ra.
- 706Ogawa S, Lee TM, Kay AR, Tank DW. 1990. Brain magnetic resonance imaging with contrast707dependent on blood oxygenation. Proc Natl Acad Sci U S A 87: 9868–9872.
- Raffy G, Mercier F, Blanchard O, Derbez M, Dassonville C, Bonvallot N, et al. 2017. Semi-volatile
 organic compounds in the air and dust of 30 French schools: a pilot study. Indoor Air 27:114–
 127; doi:10.1111/ina.12288.
- Rauh VA, Perera FP, Horton MK, Whyatt RM, Bansal R, Hao X, et al. 2012. Brain anomalies in children
 exposed prenatally to a common organophosphate pesticide. Proceedings of the National
 Academy of Sciences 109:7871–7876; doi:10.1073/pnas.1203396109.
- 714Rice D, Barone S. 2000. Critical periods of vulnerability for the developing nervous system: evidence715from humans and animal models. Environ Health Perspect 108:511–533;716doi:10.1289/ehp.00108s3511.
- Rosen ML, Sheridan MA, Sambrook KA, Peverill MR, Meltzoff AN, McLaughlin KA. 2018. The Role of
 Visual Association Cortex in Associative Memory Formation across Development. J Cogn
 Neurosci 30:365–380; doi:10.1162/jocn_a_01202.
- Ruckart P, Kakolewski Kirsten, Bove Frank J, Kaye Wendy E. 2004. Long-term neurobehavioral health
 effects of methyl parathion exposure in children in Mississippi and Ohio. Environmental
 Health Perspectives 112:46–51; doi:10.1289/ehp.6430.
- Sagiv SK, Bruno JL, Baker JM, Palzes V, Kogut K, Rauch S, et al. 2019. Prenatal exposure to organophosphate pesticides and functional neuroimaging in adolescents living in proximity to pesticide application. PNAS 201903940; doi:10.1073/pnas.1903940116.

- Silver MK, Shao J, Zhu B, Chen M, Xia Y, Kaciroti N, et al. 2017. Prenatal naled and chlorpyrifos
 exposure is associated with deficits in infant motor function in a cohort of Chinese infants.
 Environment International 106:248–256; doi:10.1016/j.envint.2017.05.015.
- Slotkin TA, Seidler FJ. 2005. The alterations in CNS serotonergic mechanisms caused by neonatal
 chlorpyrifos exposure are permanent. Developmental Brain Research 158:115–119;
 doi:10.1016/j.devbrainres.2005.06.008.
- 732Stoodley CJ, Schmahmann JD. 2009. Functional topography in the human cerebellum: A meta-733analysis of neuroimaging studies.NeuroImage 44:489–501;734doi:10.1016/j.neuroimage.2008.08.039.
- Sudakin DL, Stone DL. 2011. Dialkyl phosphates as biomarkers of organophosphates: The current
 divide between epidemiology and clinical toxicology. Clinical Toxicology 49:771–781;
 doi:10.3109/15563650.2011.624101.
- Suskauer SJ, Simmonds DJ, Fotedar S, Blankner JG, Pekar JJ, Denckla MB, et al. 2008. Functional magnetic resonance imaging evidence for abnormalities in response selection in attention deficit hyperactivity disorder: differences in activation associated with response inhibition but not habitual motor response. J Cogn Neurosci 20:478–493; doi:10.1162/jocn.2008.20032.
- Terry AV. 2012. Functional consequences of repeated organophosphate exposure: Potential non cholinergic mechanisms. Pharmacology & Therapeutics 134:355–365;
 doi:10.1016/j.pharmthera.2012.03.001.
- Whyatt R M, Barr D B. 2001. Measurement of organophosphate metabolites in postpartum
 meconium as a potential biomarker of prenatal exposure: a validation study. Environmental
 Health Perspectives 109:417–420; doi:10.1289/ehp.01109417.
- Willcutt EG, Doyle AE, Nigg JT, Faraone SV, Pennington BF. 2005. Validity of the Executive Function
 Theory of Attention-Deficit/Hyperactivity Disorder: A Meta-Analytic Review. Biological
 Psychiatry 57:1336–1346; doi:10.1016/j.biopsych.2005.02.006.
- Yolton K, Cornelius M, Ornoy A, McGough J, Makris S, Schantz S. 2014. Exposure to neurotoxicants
 and the development of attention deficit hyperactivity disorder and its related behaviors in
 childhood. Neurotoxicology and Teratology 44:30–45; doi:10.1016/j.ntt.2014.05.003.

754 <u>Table</u>

Characteristics n=95 Study population Ν % Family characteristics European origin 95 100 % Maternal characteristics 30.4 (28.1;33.5)^a Age <30 years old 43.2 % 41 30-35 years old 40 42.1 % >35 years old 14 14.7 % Body Mass Index <18.5 kg/m² 10 10.5 % 18.5-25 kg/m² 69 72.6 % ≥25 kg/m² 16 16.8 % Post-secondary school 2 years or less 50 53.2 % 46.8 % ≥ 3 years 44 Missing value 1 Occupation at inclusion 8 no 8.4 % yes 87 91.6 % Smoker before inclusion 62 65.3 % no 33 34.7 % yes High blood pressure 96.8 % 92 no 3 3.2 % yes Diabetes no 94 100 % Missing value 1 Fruit and vegetable consumption <3 portions/day 78 82.1 % ≥3 portions/day 17 17.9 % Fish consumption < twice a month 62 65.3 % 34.7 % ≥ twice a month 33 Newborn characteristics Sex Boys 41 43.2 % Girls 54 56.8 % Gestational age ≤39 weeks of amenorrhea 44 46.3 % 53.7 % >39 weeks of amenorrhea 51 3430 (3165;3800) ^a Birth weight <3000 g 12 12.6 % ≥ 3000 g 83 87.4 % Breastfeeding 2 (0;5.7) ≤ 3 months 54 56.8 % > 3 months 41 43.2 % Child characteristics (at follow-up) Parity single child 16.8 % 16 1 sibling 45.3 % 43 ≥ 2 siblings 36 37.9 % 10.8 (10.6;11.1) * Age 10 years old 66 69.5 %

Table 1: Characteristics of participants included in the study.

		Journal Pre-proof				
		11 years old	29	30.5 %		
		Educational level				
		elementary school	69	72.6 %		
		junior high school	26	27.4 %		
		Lateralization				
		right-hander	83	87.4 %		
		left-hander	12	12.6 %		
755						
756	a.	median		(Q1;	Q3)	

Table 2: Associations between dialkylphosphate levels in maternal urines during pregnancy and activation intensity related to successful inhibition during the Go/No-Go task of their 10-12 years old children (PELAGIE cohort. 2002-2017) (FWER correction at p<0.05).

Metabolite levels	Ν	Right frontal re- gion		Left frontal re- gion		Anterior cingular region		Other brain regions	
		β (95% CI)	ext.	β (95% CI)	ext.	β (95% CI)	ext.	β (95% Cl)	ext.
DAP									
Low levels	31	Ref.		Ref.		Ref.		Ref.	
Moderate	28								
levels High levels	31								
DM						X			
Low levels	38	Ref.		Ref.		Ref.		Ref.	
Moderate levels	25				•	0			
High levels	27			-1.2 (-1.6 ; - 0.8)	465				
DE				0					
Low levels	32	Ref.		Ref.		Ref.		Ref.	
Moderate levels	29	-1.8 (-2.5 ; - 1.1)	1072	-1.7 (-2.4 ; -1.0)	951				
High levels	29	-1.6 (-2.0 ; - 1.2)	390						

757 Note: DAP: Dialkylphosphate, DE: diethylphosphate, DM: dimethylphosphate

758 β (95% IC): estimated coefficient [95% confidence interval]. Empty cell: no statistically significant

759 difference reported in this region.

760 Estimates are provided for clusters presented in Figure 3. All estimates are from mixed-effect gener-

761 alized linear regression models adjusted for maternal fruit/vegetable consumption (< 3 portions a

762 day vs \ge 3 portions a day) and the duration of breastfeeding (< 3 months vs > 3 months).

763

764 Figure legends

Figure 1: Flowchart of inclusions from the PELAGIE cohort (n=3,421) to the present study (n=95). Figure I. Flowchart of inclusions from the PELAGIE cohort (n=3,421) to the present study (n=95).

765 Note: WA: weeks of amenorrhea

766

Figure 2: Associations between OP metabolites levels in maternal urine during pregnancy and performance indicators for the Go/No-Go task of their 10- to 12-year-old children (PELAGIE cohort, 2002-2017) (n =92).

- 767 Note: DAP: dialkylphosphate, DE: diethylphosphate, DM: dimethylphosphate
- 768 All estimates are from linear regression models adjusted for maternal fruit/vegetable consumption (<
- 769 3 portions a day vs \ge 3 portions a day) and the duration of breastfeeding (< 3 months vs > 3 months).
- 770 *p < 0.05, \dagger : 0.05 \leq p < 0.10.

771

Figure 3: Summary of associations between urinary OP metabolites levels during pregnancy and brain activation during successful inhibition $(n = 90)^*$.

- 772 Note: DAP: dialkylphosphate, DE: diethylphosphate, DM: dimethylphosphate, inf.: inferior, L: left, R:
- 773 right
- Adjusted for maternal fruit/vegetable consumption (< 3 portions a day vs \ge 3 portions a day) and the
- duration of breastfeeding (< 3 months vs > 3 months).
- *FWER corrected- $p \le 0.05$ with a one-sided uncorrected- $p \le 0.01$ threshold.

777





	Low level	Moderate level (vs low)	High level (vs low)
DAP	Reference	No statistically significant difference	No statistically significant difference
DM	Reference	No statistically significant difference	Superior view L. inf. frontal
DE	Reference	Lateral (left) view Superior view L. frontal R.frontal Lateral (right) view	Superior view R. frontal